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INTRODUCTION

Contemporary civilization is characterized by an exceedingly fast development of science and technology. Outstanding achievements in various fields have had as consequences the creation in many countries of large industrial centres with a production satisfying the ever increasing needs of the world population. ✓

Although amazing achievements have been obtained in contemporary industry, its rapid development stirred up greater and greater anxiety, because industrial enterprises are a source of negative phenomena menacing man's health or even endangering his life. The phenomena of increasing concern to the specialists are air pollution and surface water pollution, which appear with a particular intensity in the large urban agglomerations where is a significant concentration of industries.

More and more frequently one reads, both in the specialist literature and in newspapers, of cases of environmental pollution that result in considerable social damage. Among the most serious should be mentioned first of all the chronic and acute mass intoxications, the disturbances caused in the physical and psychical state of people living in areas with a high degree of process gas and dust contamination, the loss of strength of humans working and living in such areas, the destruction caused by the noxious materials released in air and water to fauna and flora, and the modification of the ecological equilibrium, which is sometimes irreversible.

It must be also mentioned that air and water pollution brings about not only damages of great social significance but also losses of an economic nature that can be more or less estimated in terms of money. Such losses are increasingly important in the economy of the highly industrialized countries because they cannot be ignored when the consequences caused by the influence of the negative phenomena accompanying the development of the contemporary industry are analysed. To illustrate the magnitude of such losses due to air and water pollution, it is enough to quote some figures from the world literature.

For instance, on the basis of the investigations carried out in the United Kingdom of Great Britain and Northern Ireland, it was found that the corrosion of the railways due to air pollution is now six times faster than in the early days. In France, zinc-plated sheet roofs now last about 4-5 years instead of 30 years. In Paris, Venice and many other cities, rapid destruction of the plaster of buildings and historical monuments has been reported. It has been proved that concrete not protected with a coating deteriorates quickly under the influence of sulphur, chlorine, nitrogen, carbon compounds; after only a few months in service, holes can easily be made in the concrete. In Poland, according to a survey made in 1965, it was found that over 250,000 ha of forest had been destroyed because of air pollution. The direct losses due to the decrease in the brush growth and the additional planting expenses were estimated at over 300 million zlotys annually.

The combined effect of several types of noxious materials may lead to the destruction of flora over large areas, to soil erosion and finally to landslides, with incalculable social and economic consequences. Severe air pollution leads to the increase of light absorption in the atmosphere, resulting in an increase of the energy consumed for lighting. It has been calculated that in Leningrad the losses due to the effect represent about 28 per cent of the total energy consumed in the city.

In the case of water pollution, specialists have calculated that about \$40 billion would be required to clean up Lake Erie in the United States of America.

The above-mentioned figures representing the losses caused by air and water pollution are only examples and do not give the complete picture of the negative economic consequences accompanying contemporary industrial development. In order to illustrate this more accurately, it is necessary to have data by means of which one may calculate the economic losses produced by environmental pollution as a result of the operation of improperly equipped manufacturing plants. Similar data on air pollution have been established on the basis of the investigations carried out in the United Kingdom and the United States. In the United States, the calculation shows that 1 t of impurities in the air caused losses representing an average of \$400 (the losses are estimated only for agriculture and urban areas without including the losses in the industry and the transportation sector).

were estimated at \$270/t. If it is taken into account that, according to the same calculation, about 145 million tons of different impurities are annually released into the atmosphere in the United States, one may estimate that annually this country records losses up to about \$60 billion attributable to air pollution alone. According to the calculations made, in order to provide tolerable life conditions the United States would have to spend for atmosphere protection equipment alone an amount of about \$300 billion, i.e., about \$10 billion per year from 1970 to 2000. In fact, that was the sum requested for environment protection by the President of the United States in his "State-of-the Union" message in 1970.

Although the examples presented here mostly refer to atmospheric pollution, water pollution is just as serious a problem. It is unnecessary, however, to prolong this introduction; the examples given here are sufficient to show the social and economic losses that may be caused by air and water pollution when plants are not provided with facilities for collecting and destroying the noxious matters they discharge into the environment.

This paper is intended as a concise analysis of the impact of the petrochemical industry on the environment. It identifies the pollutants produced, describes their effects on the environment, specifies the maximum allowable concentrations of these noxious materials, and discusses the treatment methods necessary to keep the petrochemical contribution below those limits. The final chapter gives two concrete, detailed examples of water-pollution control in petrochemical complexes.

It should be emphasized that this paper deals with the reduction of the pollution produced only by petrochemical processing plants and their auxiliaries. That pollution represents only a thousandth of the pollution brought about by the use of the petrochemical products themselves. More attention must be directed towards the reasonable use of these products.

For example, it was pointed out at the United Nations Conference on the Human Environment, held at Stockholm in 1970, that the application of fertilizers and pesticides was not being carried out reasonably and scientifically, leading to the pollution of soil and ground water, the destruction of fauna over large areas and the contamination of surface water. Another example

is the unreasonable use of plastic materials. According to the specialists, it is possible that sometime during the period 1980-1985, the production of plastics will become equal to that of steel.^{2/} If in the case of ferrous and non-ferrous metals a recovery and reuse system has been established in almost all countries, an analogous action in the plastic field is still in its initial stages and this only in some countries (Federal Republic of Germany, Japan, Great Britain, United States).

Today, there are several manufacturing processes for almost every product. Specialists have to select the process with the least pollution tendency or for which the most effective, not necessarily the most economical, treatment plants can be utilized. If that principle is ignored negative effects with even more serious economic implications may appear later.

^{2/} "Trends and concerns in the plastic material industry", Plastic Materials Review, No.1/1974, pp. 39-41.

1. THE IMPACT OF THE PETROCHEMICAL INDUSTRY ON THE ENVIRONMENT

One of the industries that contribute the most to environmental pollution is the chemical industry. Among the various branches of that industry, petrochemicals is considered as one of the worst offenders owing to the huge amount of chemicals that pass through the plants, whether they are simple refineries or large complexes for the cracking and petrochemical processing of the intermediates, and the high toxicity of the products. It is natural that, from the high volume of materials processed and in processing, some particles that are quite minute, and from the erosion standpoint but rather high in absolute value will be lost during the processing either in atmosphere or in the waste waters.

The restrictive measures concerning poisonous discharges in the environment imposed during the last two decades upon industry have led to an increase of investments for environmental protection within the chemical and petrochemical industries. The amounts spent by the chemical companies in the United States to reducing pollution were 4.7 per cent of investment in 1970, 8.2 per cent in 1971 and 12.1 per cent in 1972. For 1973, all the 137 chemical companies members of the Manufacturing Chemical Association have reported that they would spend \$1.4 billion for this purpose. The average investment cost for environmental protection will increase to 9.2 per cent in Japan, 9.9 per cent in the United States and 6.1 per cent in Sweden. All these financial efforts are meant to continuously reduce the noxious materials released into the air, water and soil.

Air Pollution

There is the tendency to consider air pollution as a recent phenomenon. But the problem was already present in the early days of the Industrial Revolution and has since become more and more prominent. As far back as the second half of the 19th century, the problem was severe in London, but the citizens' protests were dissipated in the rush of industrial development.

✓ Chemical Week, vol. 110, No. 21 (May 1973), pp. 59-64.

✓ Bulletin d'Informations Neurtay, No. 60 (March 1974), p. 11.

In the United States, legislation to control air pollution appeared in 1863 in Massachusetts.

The effects of pollution on man's health may be serious. Cases of collective intoxication are known, such as that in 1948 at Donora, Pennsylvania, United States, when almost half of the town population (about 14,000) were either ill or died. But the greatest disaster took place in London in 1952 when mortality increased by 1,600 per cent because of the so-called "killing smog".

Along with respiratory pollution, chronic diseases have appeared, such as emphysema, bronchitis, cardiovascular diseases, pulmonary cancer and so on. The individual effect of each pollutant are not yet known, but the effect of general retention of pollutants is well established.

From published data of recent years, specialists have established approximately that the total amount of release of noxious gas and dust in the total atmosphere is up to 40,000 million t/a, of which about 150 million is sulphur dioxide. If the data mentioned above are applied to these values, we may estimate the amount of losses caused by air pollution on a world level - 400 - 500 million.

The contribution of the petrochemical industry to these losses is substantial due to the extremely toxic materials that are exhausted in large amounts. Among these are saturated and unsaturated, closed-ring and open-chain, hydrocarbons; monofunctional compounds (alcohols, phenols, ethers, aldehydes); difunctional compounds (aldehydes, ketones); carbon, nitrogen and sulphur oxides; hydrogen sulphide, aerosols; and powders.

The amount of pollution depends strongly on atmospheric conditions. In certain regions and times of year, the winds and the thickness of the atmospheric layer may contribute to the dispersion of large amounts of pollutants; when the atmosphere is stable, much lower quantities of toxic gas accumulate in a given concentration. The nature of the terrain also has an effect: Places situated in natural basins, at river confluences, on gulfs, or on the shelter of mountains are more exposed to pollution because of the temperature inversions, which stabilize the air, are more prevalent there.

Surface-water pollution

Surface-water pollution represents one of the most important of the effects on the environment caused by the petrochemical industry. Certain groups of pollutants, including inorganic acids, bases and salts, all types of hydrocarbons, organic compounds with one to four functional groups, and various suspensions of dimers, trimers and polymers, are discharged by the large petrochemical units.

If the introduction stressed the destructive action of the toxic materials spread into the atmosphere, this subsection will dwell upon the over-all effects of the above-mentioned groups of products on aquatic ecosystems.

The evacuated pollutants, some of which are highly poisonous, strongly affect aquatic life even in quite low concentrations. Thus, hydrogen sulphide, ammonia, free chlorine, phenols, amines, cyanides, unsaturated hydrocarbons in low concentrations may change the organoleptic, physical, chemical and biological characteristics of water, giving it a repugnant taste and odour. The same effects are also given by alcohols, esters, ketones, aldehydes, organic acids and oil products in low dosage.

The action of these toxic materials on aquatic life is complex. The various organisms populating an aquatic basin and the abiotic medium are compulsorily interdependent. The connexions between them are achieved in various ways, mainly by means of the trophic element, and confer an integrality to the entire biocenosis. The integrality represents the result of the structural and functional differentiation of the species forming the biocenosis, manifesting itself by a more or less advanced mutual adaptation. Its development direction is that of the optimum use of the materials and power resources proper to the given ecosystem.

Pollution of a water basin, i.e., the introduction of certain chemicals, influences the integrality of the respective biocenosis and upsets its equilibrium. Contaminant biodegradable organic substances produce such disturbances rather by their amount than by their nature. Presumably, the aquatic organism's living in the basin are already adapted to the presence of such materials, in the form of dead plants and animals, and are able to degrade and mineralize them. But their large amount effects some biocenotic modifications: an increasing density of the species directly using the organic materials as a nutritive substratum, the appearance of

others that suffer the loss of the oxygen consumed in the metabolic process, and the disappearance of species characterized by an increased requirement of oxygen. Hence, a reorganization of the biocenosis accompanied by the alteration of its integrality is produced. The degree and rate of displacement and of integrality decrease depend, on one hand, on the kind and concentration of the injurious acid, on the other hand, on the sensitivity of the breeding species, each vegetable or animal species presenting a certain degree of resistance that amounts to a certain tolerance level.

With inferior organisms, the toxic materials penetrate through the whole surface of its body. As they differentiate morpho-physiologically, special entry passages are penetrated, for example, the digestive system and the breathing apparatus by means of which the toxic materials come into the circulatory system, which in its turn carries them to various organs and tissues. When the toxic materials reach the cell level, they act on the cell metabolism in the various ways according to their chemical composition and depending on the chemical composition of the cell substratum. So, they may influence the permeability of the cell membranes, producing disturbances in the osmoregulation mechanism; act on the enzyme equipment, inhibiting some enzyme activity; or react with certain radicals of the proteinic substances, modifying the physico-chemical composition and the aggregation state of the protoplasm.

Various toxic materials may selectively act upon certain parts of the organism, but, because of the high integrality of the living matter at the organization level of the individual, the action of the toxins on one part also has repercussions on other parts with which it is in a close functional relationship.

The living organism, which is a self-control system, tends to counteract the poisonous action of the toxic materials, i.e., preserve its integrality by means of various mechanisms. The reactions of the organism in this respect vary with the nature of the toxic material, corresponding to the action of the latter.

For instance, fish defend themselves against the waterborne action by accumulating a certain amount of mercury; when the accumulated amount reaches a certain level, the response varies. Since they affect the rate of the metabolic processes, the main reaction is such that the metal is stored in the liver and the elimination from the body of the metal is inhibited. In the modification of the environment systems may occur, with the toxicological process the toxic products tend to convert the toxic material into a non-toxic form (e.g., conversion of toxic DDT to benign dieldrin or to dieldrin (BDE)).

A particularly important aspect of the toxicity of toxic materials is their accumulation in certain tissues of the body. The heavy metals, lead, iron, copper, zinc, such as lead, liver, bones. Some substances, such as chlorinated hydrocarbons, oil hydrocarbons, etc., deposit especially in the fatty tissues, becoming harmful for a while. Under certain conditions, for example, when the organism is subjected to a starvation period, the toxic materials present in the circulation, produced at some internal organs, are released.

Finally, the toxicity of a material, i.e., the intensity and rapidity with which it acts, depends on two types of factors: those relating to the environment where the organism is integrated - the external or ecological factors - and those characteristic of the organism upon which the action is performed - the internal factors.

The external or ecological factors influence the toxicity of a material in two ways: by changes in the intensity of the metabolic processes and by modification of the physical and chemical properties of the toxic material. Generally, it is difficult to separate these two categories of effects because they usually appear at the same time.

The main external factors influencing the action of some toxic materials may be the following: the interaction between the toxic materials, the environment temperature, the amount of oxygen dissolved in the water, the CO_2 content, pH, the mineral salt content, light, turbidity.

Soil and ground-water pollution

The toxic products that may pollute the soil and, by leaching, also the ground-water layers are practically the same as those encountered in surface-water pollution. Their presence on the soil at undesired places and in the ground-water layer is due to the following main causes:

(c) the presence of cracks in the water lines in the case of manual input of the water into the water works areas if some water that can corrode the water pipes at such the sewage line are used;

(d) the depth of penetration of the soil or an inadequate drainage of some water due to partial leaching of the products into the ground-water layer in the case of the waste of the wastes or the increase of humidity of the soil;

(e) the presence of cracks in the soil or after a process failure and water entering into the refuse spouts.

Finally, one or all the above may lead to the soil contamination of the inside of the area of the plants with direct consequences on the ground-water layer.

Because of its location into the ground water, it does not purify itself as readily as the surface water, and therefore the effect has a more persistent character. Finally, these layers are used both for the water supply in areas that are short of surface water and for the water supply in the industrial units. Therefore, the effects of the aquatic layer contamination will have direct consequences, both social and economic. Also, it is necessary to keep in mind that there is a continuous communication between the ground-water layers and the surface waters, so that pollution of the ground-water layer will sooner or later lead to pollution of the surface water with which it communicates.

NOISE

The main sources of noise in petrochemical complexes are the pump and compressor units and the high-pressure thermal equipment in the nitrogenous fertilizer plants. There is nothing unique in this aspect of pollution by petrochemical industry. A brief discussion of the salient points is given in chapter IX.

11. ALLOWABLE POLLUTANT LEVELS

At the outset, it should be stated that the establishment of maximum allowable levels (limit values) for pollutants within a given country is a matter of regulation and depends on the country's environmental protection policy. For border regions and basins of water shared by several countries, regulations are established by joint commissions taking into account the principal international standards. Therefore, the limit values for air and water pollutants given here are only indicative in character. It is recommended, however, that countries which have not yet legislated maximum allowable levels consult the pertinent ASTM Standards of the United States, the GOST Standards of the Union of Soviet Socialist Republics, the British Standards of the United Kingdom and the International Standards of the International Standards Office.

Air pollutants

Annex I is a table giving the limit values established in the United States and the Union of Soviet Socialist Republics for some 400 inorganic materials that can be air pollutants.

The United States limit values were established by the American Conference of Government Industrial Hygienists (ACGIH)^{5/} and are characterized as follows:

- (a) They are experimental values obtained in industrial practice and in experiments conducted on animals;
- (b) They are valid for momentary contaminations;
- (c) The determination methods differ from one material to the other. For certain materials, the limit is determined on the basis of subjective considerations, for others, it is based on observations characterized by short or long term interruptions. Hence, these values do not have the same significance for all the materials, and no distinction can be made between dangerous and non-dangerous materials;

^{5/} "Threshold Limit Values for 1972" (Cincinnati, ACGIH).

(3) The above-mentioned value is defined as the concentration to which workers can be exposed daily without unfavourable consequences. They are average values for a working schedule of seven to eight hours a day, five days a week.

A limit value marked with the letter P in column 5 is the threshold limit of the concentration. A concentration of this magnitude does not present any danger, but if it is exceeded for 15 minutes, the following may occur:

- Intolerable irritation
- Chronic or systematic modification of the tissues
- Inhibition of self-preservation reflexes
- Reduction of working efficiency

The following relation between the units will help in interpreting the table (at temperature of 20°C and pressure of 760 Torr is assumed):
concentration (mg/m³) = concentration (ppm) x molecular weight/24.45.

The Soviet limit values were established by Public Health Ministry Standard No. 45-11 on November 1971 and are enforced by the State Sanitary Inspector. They are defined as the concentrations at which no occupational disease occurs during an ordinary eight-hour working day.

In both the United States and Union of Soviet Socialist Republics standards, certain materials (marked by an "x" in columns 5 and 7 of the table) are said to have "striking toxicity", i.e., they cause intoxication not only via the respiratory and alimentary systems but also through the skin.

Surface-water pollutants

Some countries use several categories of standards when they establish the limits for surface water pollutants. For example, in Romania surface-water supplies are divided into three categories according to and use as follows:

Category I

- Drinking
- Industrial units
- Animal breeding units
- Food and other industries requiring drinking water

Certain irrigated agricultural crops
Salmonoids fish culture
Swimming pools and basins

Category II

Fish culture except salmonoids
Breeding and development of native fish in natural bodies of water
Certain industrial units
Urban and recreational use

Category III

Irrigation
Industrial processes

Annex II is a table giving the quality standards for the waters in these three categories. The data are according to Romanian Standard No. 4706/1966.

Besides quality standards for surface water, there are also quality standards for effluents. These may be established separately for each industry (e.g., oil processing, petrochemical, fertilizer, dyestuff, sugar) or for certain sections of a body of water on which industrial units are located, e.g., a case of a river, the upstream and downstream contribution from other manufacturing plants, is taken into account when imposing specific quality indexes on a given industrial unit.

In establishing these indexes, the amount of pollution from upstream, the surface-water flow, the self-purification capability of the river, the downstream requirements and other limitations are taken into account. That is why the limitations imposed in some cases may be more severe than those stipulated in the regulations concerning the quality of the water supply.

As these quality standards differ when the same river passes from one country to another, it is difficult to make recommendations in this field. Such data are given in the literature, but without any comment on the factors determining the establishment of the limit values.

On the other hand, there are cases when the limit value is not to be set for the pollutant concentration in the effluent to a certain extent stipulated in the permit on water quality. This occurs in water courses with high flow velocity (with pumping power), where the self-purification process takes place satisfactorily, and where the pollutant density in certain areas is very low, meaning that the distance between two waste-water discharges, or the industrial or urban, is relatively high, say more than 50 km.

The condition that must be met for each pollutant if the waste water containing it is to be discharged into the river at a different location, without preliminary treatment is

$$\sum_{i=1}^n q_i C_i < Q_{uz} C_{ad}$$

where

- q_i (i=1, ..., n) = Average rate of discharge of waste water at location (in l/s or m³/h)
- C_i (i=1, ..., n) = Average concentration of the pollutant in the waste water discharged at location (in mg/l or g/l)
- Q_{uz} = Average waste-water flow rate in the river (in l/s or m³/h)
- C_{ad} = Maximum permissible average concentration of pollutant in the river (in mg/l or g/m³)

Note that the contribution of pollutant to the effluent, the product $Q_{uz} C_{ad}$ above, has the dimensions of a rate (mg/s or g/h).

Soil and ground-water pollutants

The specialized literature has very little data on contaminants discharged onto the soil. Practically speaking, the maximum allowable concentrations of various materials and groups of materials depend on the flora and fauna surrounding the plant. Usually, these limits are established by agricultural experts and silviculturists.

For ground water, the maximum allowable concentrations of pollutants in most countries are the limit values corresponding to category I of surface water.

Noise

In almost every country, the maximum allowable level of the "pollutant" noise has been set in the range 90-100 dB. (See chapter IX).

Annual reports of the International Commission on Environmental Health (I.C.E.H.) have shown that particulate matter, sulphur dioxide, nitrogen dioxide, carbon monoxide, and lead are the most important pollutants from the combustion of fossil fuels. The I.C.E.H. has also shown that sulphur dioxide, nitrogen dioxide, and carbon monoxide are the most important pollutants from the combustion of fossil fuels in the industrial sector.

Particulate Matter

The principal source of air pollution from manufacturing and combustion processes is the supply power for the manufacturing processes and the manufacturing processes themselves. The main pollutants are sulphur dioxide, sulphur trioxide, hydrogen sulphide and carbon monoxide.

Carbon monoxide appears as a result of incomplete oxidation reactions in either manufacturing or combustion processes. Its negative effects are due to its combination with the haemoglobin in the blood of mammals, producing highly stable carboxyhaemoglobin. The consequences are various: vascular and pulmonary diseases, anaemia and aggravation of thyroid diseases. Studies show that breathing air containing as little as 1 ppm carbon monoxide is harmful to mental activity.

Sulphur oxides appear as a result of the complete oxidation of hydrogen carbons contained with sulphur compounds. If the sulphur content of the crude oil to be processed is above certain limits, the oil is desulphurized. Nevertheless, the sulphur that remains in the feedstock, or in a part in the fuel the plant uses will still appear in the waste gas in the form of sulphur dioxide or trioxide.

Sulphur oxides irritate the upper respiratory tract and, together with any solid pollutant particles that get into the lungs, are quite harmful. In the presence of oxygen and moisture, sulphur trioxide and sulphuric acid are formed in the alveoli. An annual average concentration of only 0.04 ppm physically overworks the individual, and in the presence of a concentration

See also P. Gilles, "La Pollution de la chimie face aux problèmes de l'environnement après la Conférence de Stockholm", paper presented at the International Conference on Environmental Engineering Problems in Industrial and Developing Countries, New York, September 1973 (sponsored by World Federation of Engineering Organizations, London).

of the population of the industrial zone, mortality due to bronchitis and other diseases is significantly higher. Besides having these effects on people, sulfur dioxide is the main cause of the corrosion of machinery and equipment in the industrial areas and populated centres.

The degree of corrosion of sulfur dioxide is enhanced by a high content of sulfur dioxide in the air, especially carbon dust, which facilitates the formation of sulfuric acid mist, which in combination with water forms sulfuric acid droplets, respectively. That explains why the vegetation in the industrial zone and sulfur dioxide mist content is virtually completely destroyed. In such areas one can see a great deal of erosion and high incidence of occupational diseases. Also, in these areas, industrial equipment and buildings have to be replaced from time to time as often as in other areas.

Hydrogen sulfide is a by-product of hydrorefining and desulfurization. It is a toxic gas, acting especially on the nervous system. In concentrations above 10 ppm in air for long periods, it may cause death. It has the same effect on warm-blooded animals. Because of its strong and specific odor, people do not work well in places polluted by it.

Hydrocarbons escape into the air during decantation and handling of various types of finished, intermediate and raw products, mostly through leaks in the processing and transport equipment. Generally speaking, these pollutants are less toxic than those mentioned above. The danger from them arises from the fact that they are a major factor in photochemical fog formation. This recently discovered phenomenon^{2/} contains photochemical oxidants that directly affect health. At a concentration of 0.1 ppm, eye irritations occur and at 0.05-0.06 ppm concentrations, the frequency of asthma crises is increased. The adverse effects of these products is being studied, the existing data being insufficient to draw detailed conclusions about them.

^{2/} Annual Report of the Council for Environmental Quality presented to the Congress of the United States, August 1970.

Among the economic effects of these contaminants an important one is that the air inlets of oxygen-manufacturing plants (hydrocarbon-polluted areas) must be situated a great distance away from non-polluted regions to avoid the danger of explosion.

Nitrogenous fertilizer plants:

The main contaminants appearing during nitrogenous fertilizer production are carbon oxides, nitrogen oxides, ammonia, sulphur oxides in case the methane gas contains sulphur compounds (mercaptans), and dusts, e.g., ammonium nitrate dust.

The nitrogen oxides appear in high concentrations only in the production of nitric acid by ammonia oxidation. The noxious effects of the nitrogen oxides are due partly to nitrous and nitric acid formation in the atmosphere in the presence of moisture, and partly to their role, in sunlight, as initiators of chain reactions involving gaseous hydrocarbons which produce a range of secondary pollutants called photochemical oxidants. These oxidants, together with the solid and liquid particles in the air, form the so-called "smog" and cause irritation of the mucous membranes and the respiratory tract.

The pollutants from the photochemical oxidant family also include ozone (an unstable, toxic form of oxygen), nitrates, peroxides, aldehydes, acrolein etc.

Concerning the toxic effects of nitrogen oxides on man, fauna and flora, little is known aside from their contribution to smog formation. One fact is that small amounts increase a person's susceptibility to Asiatic influenza.

The carbon oxides appear in very high quantities in synthesis gas production, in the methane gas cracking unit.

Ammonia appears in all the steps of the nitrogenous fertilizer production (nitric acid, ammonium nitrate and urea preparation). It is an irritant to mucous membranes and the respiratory tract. If sufficient quantities are breathed for a long time, it produces maladies of the digestive and respiratory systems. It may also produce eutrophication effects on the flora.

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One of the main negative effects of the entire above-mentioned

Auxiliary Plants

For technical and economic reasons, petrochemical complexes include, besides the basic plants, auxiliary plants for upgrading existing raw materials or for supplying the additional feedstock required for

Examples of plants used for the production of carbon black, with an amount of particulate emissions. The contaminants released by these plants are carbon black and particulate matter, which are released into the atmosphere through leakage.

Carbon black, by itself, is a pollutant only from the very moment of production point of view. However, in the presence of sulphur and nitrogen oxides it facilitates, because of its high absorptivity, may form and accelerate action of these oxides, contributing at the same time to smog formations.

Particulates, of course, are intended for the controlled and selective destruction of a desirable plant and insect organisms. They are dangerous to other life when allowed to escape into the air in any form or an uncontrolled manner. It should be pointed out that some of these plants do have a cumulative effect and they may cause irritation either by direct contact with the product or by ingestion of food contaminated by it (DDT is a well-known example).

2. EFFECTS OF TOXIC MATERIALS ON AQUATIC ORGANISMS AND THE CHEMICAL EQUILIBRIUM OF THEIR RESERVES

Noxious materials present in the industrial effluents contain several toxic substances. The toxicity of these materials varies according to their nature: (1) synergism, (2) additive (cumulative), (3) greater than additive (synergism), (4) lower (antagonism).

The physical characteristics of the surface water, such as its temperature, amount of dissolved oxygen, salinity, and other factors, amount of suspended matter, etc., have a considerable effect on the action of the poisons on aquatic plants and fauna.

In general, the toxicity of poisons to aquatic organisms according to their effect is: (1) the rate of increase in O_2 , within certain limits or at a certain level, (2) the rate of metabolism and with it the rate of production of toxic products. Temperature also affects the physical and chemical characteristics of the water, gas solubility, electrolytic dissociation, etc. In general, a temperature increase causes an increase in the toxicity of the poisons to organisms.

Dissolved oxygen also influences the metabolic processes of aquatic organisms and their resistance to toxic materials. For example, the action of a toxic material is not so severe in water with a low dissolved-oxygen content, resulting in a decrease in the fish survival time, or in other words an increase in toxicity. The explanation is that many noxious materials act primarily on the breathing apparatus, reducing its efficiency. In oxygen-saturated water, the oxygen produced by the organism is consequently reduced, resulting in a lowered chance for survival.

A high concentration of dissolved oxygen in the water, on the other hand, accelerates the oxidation of any oxidable noxious material present, leading to detoxification of the environment and a prolongation of organism survival time.

In the case of oxygen-saturated water, the response time of the organism to the noxious material is inversely proportional to its oxygen consumption, namely the intensity of its breathing. Thus, the trout, which consumes oxygen at a high rate, responds more quickly to the toxic action of a material; therefore, it is more sensitive than, for example, the crucian, carp and tench, which do not need so much oxygen.

The carbon dioxide content directly influences the respiration of aquatic organisms. Large amounts of carbon dioxide in water containing toxic materials, generally accelerate intoxication. High CO_2 concentrations in water make it more difficult for the CO_2 produced in respiration to diffuse through the bronchial membranes, leading to its accumulation in the body. This accumulation causes the blood pH to decrease with serious effects on the state of the organism. Also, the higher the carbon dioxide concentration in blood, the more difficult the combination of haemoglobin with dissolved oxygen. Therefore, the presence of large amount of CO_2 increases the minimum limit of the oxygen concentration.

The carbon dioxide content may also indirectly influence the noxious effect of some impurities by its effect upon environmental factors, especially the pH. It is known that the toxicity of a material may depend upon its dissociation rate in water and this is determined by the pH. For example, ammonia is more toxic in water with high pH (alkaline), where it largely retains its identity as NH_3 , than in water with low pH (acid), where it reacts to form the ammonium ion NH_4^+ , which is less toxic than NH_3 .

The cyanides and sulfides, on the other hand, are more toxic in an acid medium than in an alkaline one. Fish may tolerate a concentration of cyanides combined with nickel which is 1,000 times higher at pH 8 than at pH 5. A decrease of 0.3 in pH in a nickel sulphate solution leads to a ten-fold increase of toxicity.

The mineral content of water greatly influences the toxicity of noxious materials. Generally, the toxicity is lower in hard water than in soft water. The decrease in toxicity is due primarily to chemical reactions that result in the precipitation of some non-toxic salts and secondarily to alterations in cell-membrane permeability.

Light has a direct action on some noxious materials. Some cyanides are decomposed by light with the release of highly toxic cyanide ions. The action of light may also create toxic materials from some tars.

The toxicity of some materials is much decreased in highly turbid water owing to their being adsorbed on the suspended particles.

Petroleum refineries

Effluents from oil refineries contain mainly the following contaminant products or groups of products: mineral acids, sodium hydroxide, sulphides, phenols and cyanides.

Mineral acids may appear in the waste water either from the manufacturing processes or from the waste water treatment. They may have a direct destructive effect on flora and fauna or an indirect effect by alteration of pH.

Sodium cyanide results from the desulphurization processes employed in refineries. It is a corrosive poison that directly affects fish branchia and aquatic flora.

Sulphides are potential reducing agents and have high toxicity even in small concentrations. At pH 9, 3-4 mg/l sulphide represents the lethal dose for most species of fish. The toxicity increases with decreasing pH, e.g., at pH 7 the lethal dose for fish is 1 mg/l (see foot-note on page 25). The sulphides are quite toxic for the microfauna that serves as fish food; it is practically destroyed at sulphide concentrations of 10 mg/l.

Phenols are not always present in the waste water from the refineries, but in some cases they can attain very high concentrations. The phenols are usually carried off by hydrocarbons during cycle failures or jet overflows. The phenol concentration is variable depending on the origin of the crude oil and the characteristics of the final products. The main negative consequences they have on water are the following:

(a) They are very toxic for fish and microfauna. At 1-2 mg/l concentration, the micro-organisms are destroyed; at 5-10 mg/l, the salmonoids are destroyed; and at 10-15 mg/l, the cyprinids are destroyed;

(b) At 0.1 mg/l concentration, the flesh of food fish is no longer edible;

(c) The phenol presence is quite troublesome in the treated water used for drinking water supplies because in combination with chlorine or hypochlorite (both used instead of ozone as oxidants for organic materials) they give chlorophenol, which is perceptible even in concentrations as low as 1 µg/l.

Hydrocarbons appear in the waste water from the manufacturing process by being carried off by rainwater from the industrial platforms or accidentally. Their toxic and destructive action on fauna and flora is evidenced in these ways:

- (a) A decrease of biodegradability, which can lead to increased accumulation that hampers vital exchanges and impedes the development of fauna and flora;
- (b) The carcinogenic effect of some hydrocarbons, which by degradation release benzopyrenes;
- (c) Cumulative phenomena caused by chain concentration in food;
- (d) A decrease of oxygen exchange between air and water because of a continuous film formed on the water surface;
- (e) High toxicity for aquatic organisms. The harmfulness varies depending on the constituents, e.g., for benzene and toluene the limit is established between 10 and 90 ppm and for naphthalene and anthraquinone hydrocarbons, the limits are as low as 4-7 ppm. The paraffin hydrocarbons are less noxious but they create biological derangements (fauna migrations); similar phenomena are experienced with the olefin hydrocarbons. The hydrocarbon fractions with molecular (sulphur, metal compounds etc.) have a toxicity comparable to the aromatics.

Nitrogenous fertilizer plants

The effluents from nitrogenous fertilizer plants contain nitrogen compounds such as: nitric acid (HNO_3), ammonium hydroxide (NH_4OH), ammonium nitrate (NH_4NO_3) and urea ($\text{CO}(\text{NH}_2)_2$) coming from the nitric acid, ammonia, ammonium nitrate and urea plants, respectively.

The main effect of all these products is that by dissociation they release nitrate (NO_3^-), ammonium (NH_4^+) ions and CO_2 , which lead to the eutrophication of bodies of water. Ammonia attacks the breathing systems of fish, more or less severely, depending on pH and temperature. An NH_3 concentration of up to 0.025 mg/l is tolerable by fish. The table gives the total ammonia + ammonium-ion ($\text{NH}_3 + \text{NH}_4^+$) concentration that yields an NH_3 concentration of 0.025 mg/l as a function of temperature and pH.

1.0 mg/l of nitrate (NO₃) is added to 1.0 ml of water, the concentration is 0.001 mg/l

Time in %	Concentration					
	1	2	3	4	5	6
5	1.1	1.2	1.3	1.4	1.5	0.088
10	1.2	1.3	1.4	1.5	1.6	0.068
15	1.3	1.4	1.5	1.6	1.7	0.054
20	1.4	1.5	1.6	1.7	1.8	0.044
25	1.5	1.6	1.7	1.8	1.9	0.039
30	1.6	1.7	1.8	1.9	2.0	0.035

Much larger quantities of the noxious materials mentioned above may appear in surface waters because of incorrect methods of soil fertilizer application. Some of the soluble products, nitrogenous fertilizers may be carried off by rain water into the rivers. An excessive concentration of nitrates in drinking water may cause methemoglobinemia and cyanosis in children.

Plants growing in water and other primary petrochemical products

The principal pollutants discharged in waste water by these plants are mineral acids, phenols, cyano compounds, hydrocarbons, various chlorinated products, ketones, aldehydes, heavy-metal ions, suspensions of various kinds, detergents and heat. Since the effects of the mineral acids, phenols, hydrocarbons and heat has already been discussed, this section will be limited to brief discussion of the other contaminants.

The cyano compounds, such as acrylonitrile, hydrocyanic acid, cyanogen chloride and sodium cyanide, come from plants manufacturing acrylonitrile from propylene and ammonia and plants adjacent to them. These products have a high toxicity in water, not only for aquatic fauna but also for any terrestrial fauna exposed to the water. At 50 µg/l concentration they cause the death of trout in a few days, and a concentration of 0.1 mg/l is the lethal dose for about half of the fish species. This concentration also leads to biological derangements of the aquatic environment. Toxicity increases with decreasing pH and is quite persistent.

The chlorinated products come from the vinyl chloride plants and their satellite to the petrochemical complexes. In general, the products are not biodegradable or have a low biodegradability. They are highly persistent and directly affect the fauna and the aquatic biocenosis equilibrium. At the same time they impart unpleasant odors to the water even in low concentrations, thus making the water unusable for various purposes. At higher concentrations, they cause trouble in water softening units.

The heavy-metal ions result either from the water used to purge the cooling systems or from processes in which they have been used as reaction catalysts. These ions act mainly as inhibitors of biological treatment and as poisons to fish. They also act together synergistically, or additively. For example, equal quantities of copper and zinc ions in water more than doubles the toxic action on trout and roach, whereas the combination of cadmium and zinc produces only an additive toxic action on the same species. Chromium and nickel ions also act additively.

The suspensions mentioned above consist of waste matter coming from the chemical washing of plants and their platforms and polymers having a low polymerization rate that have been carelessly discharged from equipment.

Among the undesirable properties of these suspensions are the following:

- (a) They become an impediment to the development of bottom flora;
- (b) They produce fermentation during which noxious materials are released;
- (c) They cause asphyxiation of fish due to noxious materials released either by fermentation or by contact with water;
- (d) They increase turbidity;
- (e) They impede the breathing of fish by causing deposits on the bronchia and other parts of the respiratory system.

The detergents appear in waste water as a result of industrial platform washing. Their toxic action may be direct, by their mere presence, or indirect, by creating a film on the water surface that traps suspensions and hydrocarbons. The consequences are the following.^{6/}

^{6/} "Techniques pétrolières et pétrochimiques modernes, leurs effets sur la production d'eaux résiduelles", paper presented by the French Government at a seminar in Warsaw, September 1973.

(a) Reduction of the oxygenation capability of the river, due to the surfactant effect;

(b) Destruction of aquatic herbs at 2.5 mg/l concentration;

(c) Mortality of fish at concentrations above 1 mg/l;

(d) Increased maintenance costs due to necessity of more frequent dechlorination and demineralization, etc.

Auxiliary Plants

As in the discussion of air pollution from auxiliary plants (see above), the discussion here will be limited to the contaminants originating in the production of carbon black, pesticides and chlorine by electrolysis.

In the residual water from these processes may be found carbon black, oil products, pesticides, and chlorinated hydrocarbons, mercury or mercury salts, and caustic soda and sodium hydroxide.

Carbon black, in particular, has had effects on flora and fauna that are similar to those of detergents, with the difference that it does not ferment. If the carbon black is not wettable (has a very high specific surface), the noxious effect in surface waters is similar to that of detergent foam. Because of these effects, neither form of carbon black should be used in a biological treatment plant, although the material is actually not toxic.

The mercury and mercury salts come from the electrolytic cells used in chlorine production, which have mercury cathodes. They are highly toxic for aquatic flora and fauna. The mercury finds its way into the flesh of fish in trace amounts in the food of whatever animal eats it. As with other heavy metals, mercury poisoning is cumulative. Where fish consumption is high, mass diseases may occur in chronic or acute form.

Although sodium chloride is not toxic, its presence at concentrations of 50-100 mg/l in surface water makes it unfit for uses such as agricultural irrigation, water supply of steam boilers (by reducing the ion-exchange capacity of demineralization equipment), drinking and recreational purposes.

V. SOIL AND GROUND-WATER POLLUTION BY PETROCHEMICAL INDUSTRIES

As already outlined, contamination of soil and ground water may arise from inadequate waste pits, cracked sewers, discharges of plant washing water or rain water, or from all these three causes together.

On the average, the petrochemical industry does not produce as much waste as industries such as soda ash, phosphorous fertilizer and NPK fertilizer manufacturing. Nevertheless, there are significant amounts of such wastes as polymers, acid tars, slops and sludges from the purification units. These wastes, excepting the polymers, should be pelletized when possible and stored in specially arranged places when their recovery is not economical, because their toxicity and high moisture content (93-94 per cent) may contaminate the ground-water layer or may constitute a public danger. Indeed, such wastes are better destroyed because of that danger. Soil polluted with acid tars takes a very long time to recover and should be taken out of the agricultural cycle in the meantime.

VI. METHODS OF AIR-POLLUTION CONTROL

Most petrochemical activities release into the air some pollutants that, as we have seen, represent the end of a manufacturing process. Depending on the processes and the type of equipments used, these pollutants are quite different; but they may be divided into two large groups according to their physical nature: (a) solid particles and (b) gases and vapours.^{9/}

Particles in air or, more generally speaking, in gas, appear in the form of dusts that settle spontaneously, or aerosols that do not, depending on the size of the particles. These particles arise mainly from combustion furnaces but are also released into the air in the process of storing and handling products. Those up to $10\ \mu\text{m}$ in diameter are most numerous; they can be as small as $0.01\ \mu\text{m}$.

It is generally agreed that there are only two ways to fight industrial air pollution efficiently. The first is to prevent the escape of waste gas at the source and the second, to treat and purify it. Nevertheless, the release of gas from tall stacks is one of the most frequent methods of disposal used today. Although the author does not consider that this so-called "protection" is effective, some data related to stack-height calculation will be presented for practical reasons.

After an analysis of various calculation methods, the Centre Interprofessionnel Technique d'Etudes de la Pollution Atmosphérique (CITEPA) selected the following method as the simplest.^{10/} For SO_2 emissions, the stack-height formula is

^{9/} Industrie et techniques, No. 229 (10 mai 1973).

^{10/} "Calculs de la hauteur des cheminées en France", paper presented at the May-June 1974 session of the Agence pour la Coopération Technique Industrielle et Economique (ACTIM) by M.M. Detrif and Bouscaren of CITEPA.

$$h_p = \sqrt{\frac{10.4}{n} \frac{R}{T}} \sqrt{\frac{q}{C_{SO_2}}}$$

- where
- h_p = stack height in metres
 - T = the temperature difference between the gas from combustion and the outside air in °C
 - R = exhausted gas flow rate in m³/h
 - C_{SO_2} = SO₂ concentration limit at soil level in mg/m³
 - q = SO₂ flow rate in kg/h

This relation is theoretically valid for any pollutant where the appropriate concentration limit at soil level C_{SO_2} for the pollutant is used.

For dust, the stack height formula is

$$h_p = \sqrt{\frac{680}{n} \frac{R}{T}} \sqrt{\frac{q}{C_d}}$$

- where
- h_p = stack height in metres
 - T = the temperature difference between the gas from combustion and the outside air in °C
 - R = exhausted gas flow rate in m³/h
 - C_d = dust concentration limit at soil level in mg/m³
 - q = dust flow rate in kg/h
 - n = number of stacks

The coefficient 680 = 2 x 340 takes into account the settlement effect of the dust particles.

The calculations and results presented here were checked in the Soviet Union over 100 installations during the construction of the stacks for seven thermal power stations, with the heights ranging from 40 to 150 m.

It should be noted that a calculation method for the chimney height that will be perfectly valid under all meteorological conditions and over the whole year. However, the method suggested by CITEFA gives correct results in cases of ordinary atmosphere diffusion, namely in 95-97 per cent of all cases. In extraordinary cases (significant temperature inversion, unusual wind regime) a more stringent calculation is necessary.

Solid-particle separation

By now, the means to prevent particle or dust emissions are quite well known and efficient. They may be classified according to their operating principles as follows:

- Mechanical separation by gravity or inertia
- Hydraulic separation by bubbling or spraying
- Filtration by bag or layered medium
- Electrostatic precipitation

The mechanical dust separators acting by gravity are the simplest. They consist of a settling chamber having a large volume through which the purged gas passes with low velocity so that dust may settle by gravity. However, they have drawbacks: large volume, low gas velocity and inability to separate dust particles smaller than 50-100 μm .

Inertia dust separators are manufactured in a great variety of types from simple baffle separators to cyclones and multicyclones. The efficiency of the latter is very high. For example, the use of multicyclone batteries in a nitrogenous fertilizer plant can result in the following separation yields:

<u>Particle size (μm)</u>	<u>Yield (%)</u>
2.5	68
5	85
10	95
25	99

The over-all yields, depending on the dust granulometry, are in the range 94-98.5 per cent. The maximum capacity of the cyclones used in their application is 25,000 m³/h. at a pressure drop of 100-120 mm. water gauge. Hydraulic dust separators use a liquid (usually water) to capture the dust and separate it from the gas. The efficiency of the equipment depends on the effectiveness of particle moistening, the velocity of the water particle in spraying, the gas particle velocity, water dispersion, and the diameter and number of the spraying nozzles.

There are various types having high efficiency depending on the nature of the dust: bubbling, spraying, venturi tube. The last type may have a high operating cost because an additional sludge treatment is required. All three types and the electrostatic dust separator are used for gas containing carbon-black in suspension.

Filtration separators use different filtering mediums depending on the suspension to be filtered. Here, too, there are a variety of types. The filtering layer may be cleaned manually or automatically by shaking and air blasting in a counter-current. Special attention must be given to the filtering material, which should be mechanically strong and resistant to chemical reagents and high temperatures.

Electrostatic precipitators (electrofilters) are based on the corona effect and consist of electrode rows through which the gas-containing dust is blown. The dust adheres to the collecting electrode and is eliminated. To increase the efficiency, some types of electrofilters operate with water mist. The equipment has a very high yield and efficiency, but they also have the disadvantage of a relatively high energy consumption.

Gas and vapour separation

While aerosol and dust removal is carried out rather easily by means of the equipment described above, waste gas and vapour treatment must resort to complex technical processes, since they have to deal with a continuous range of concentration of pollutants having quite different chemical properties. Four groups of air pollutants may be established: sulphur compounds, hydrocarbon vapours, nitrogen oxides, other polluting gases.

C. The second process consists of saturating the waste gas with water, then washing with a solution of calcium hydroxide (lime) or calcium carbonate (dolomite) which reacts with sulphur dioxide to form calcium sulphite or calcium sulphate. The sulphur dioxide is then removed by precipitation of the calcium sulphite or sulphate.

In the first process, the waste gas is saturated with water, then washed with a solution of calcium hydroxide (lime) or calcium carbonate (dolomite) which reacts with sulphur dioxide to form calcium sulphite or calcium sulphate. The sulphur dioxide is then removed by precipitation of the calcium sulphite or sulphate.

Another wet process is the catalytic oxidation of SO_2 to SO_3 and fixation of the latter on sulphuric acid. This process is called "Sulphacid" and may be carried out in one or two steps depending on the reactor type. The disadvantage of this process, require sulphuric acid, is more interesting than the commercial process, but the salts that can be recovered from the first process.

The drawback of both processes is that they require saturated with water into the atmosphere. They do have the advantage, however, of achieving a high retention field - at up to 90 per cent - using simple, robust equipment.¹¹

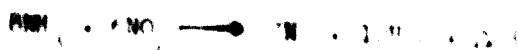
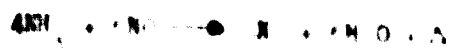
There are three important dry processes. The first involves the passage of SO_2 and SO_3 over waste products such as lime and dolomite. In this way, SO_2 is completely removed, but SO_3 is only partially removed. Another disadvantage is that the salts that are formed represent solid wastes from which sulphur cannot be recovered.

The second dry process consists of two steps: first, a catalytic oxidation of SO_2 to SO_3 , and then fixation of SO_3 on metal oxides. By a further treatment, sulphur may be recovered in the form of SO_2 , sulphates or even elementary sulphur.

The third dry process may be applied at sites where H_2S is present. In this case, a catalytic reduction of SO_2 and H_2S at about $150^\circ C$ takes place: $2SO_2 + H_2S \rightarrow 3S + H_2O + 35 \text{ kcal}$. This process is called "Sulfen" and is used in several countries, e.g., Canada and France.

¹¹ R. Joffre, "SO₂ removal from waste gas in the chemical industry", Rev. Nuisance d'environnement (October 1974).

The treatment of waste gases containing nitrogen oxides is carried out catalytically, the reaction taking place in the presence of ammonia. The waste gas is passed through a catalytic bed together with a certain amount of ammonia. The following reactions take place:



To a lesser extent, the following reactions also take place:



The process is carried out at 200-300°C. The catalyst is active below this temperature. A certain pressure, no conditions are imposed, so that the process may be applied in any nitric acid plants.

The catalyst, which is under the licence of a certain company, is not mentioned in literature. The only characteristic that is mentioned is that it is in pellet form. The amount of catalyst in the reactor must be such that the gas flow rate is at least 10,000 m³/h per cubic metre of catalyst. Under these conditions, the concentration of nitrogen oxides can be reduced to 0.1%, irrespective of their initial concentration in the waste gas.

For each reacted 1,000 ppm nitrogen oxide, the temperature increases by 14°C. For instance, if the initial nitrogen oxide content is 3,000 ppm, the gas will be 42°C hotter after the reaction.

Approximately 1.5 moles of NH₃ are required to eliminate 1 mole NO_x from the gas mixture; for 1000 Nm³/h waste gas having 1000 ppm (NO_x) content, the rate consumption of NH₃ is thus about 1.5 kg/h. The ammonia used must be gaseous, somewhat super-heated and supplied at a pressure at least 3 kgf/cm² higher than the waste-gas pressure. Additional power must be provided for the evaporation and super-heating facilities.

In practice, the reaction is operated at about 1100°C , and the ammonia is prepared in a separate plant. At the outlet, the gas is cooled, the water is separated, and then exhausted to atmosphere.

Another method of reduction of nitrogen oxides consists of mixing the waste gas with hydrogen and passing them over a metallic catalyst. The reaction is carried out at a temperature in the range of $400-600^{\circ}\text{C}$. The volume ratio of hydrogen to nitrogen oxides in the feed is about 1:1. The volume ratio of hydrogen to nitrogen oxides in the first mixture and 80-94 per cent for the second mixture, the first mixture containing 1,000-1,500 ppm NO_x , two plants are used to reduce the concentration down to 50 ppm.

The second method: the waste gas is heated and mixed with C_1-C_4 hydrocarbons and passed over a catalyst. The mixture is introduced into the reactor, the reaction is carried out carefully into a recovery oven to obtain steam, which is separated and cooled. Afterwards, the gas is released to atmosphere.

The cost of such a plant is about 10 per cent of the total investment for the nitrogen plant.

Generally, the most efficient method to treat the most efficient measures for oxide absorption in the gas and apply the catalytic reduction method only afterwards.

The purification of air containing the carbon oxides is applied only to the gas emitted from the manufacturing plants. It is not a destruction but a separation and recovery of these gases, which in the cases of nitrogenous fertilizer plants or partial methane-oxidation plants, are present in very high concentration and considerable volume. The recommended method is a separation of these gases by washing them with a hydroxide solution, which converts carbon dioxide into carbonate; the carbon monoxide that remains may be dried and used in other processes (e.g. methanol manufacture).

In the case of thermal power plants existing on the site, carbon oxide pollution can be prevented by releasing the gases into the atmosphere at a suitable height.

The key to the control of pollution caused by other products, such as ammonia, hydrocarbon vapours and cyanide vapours is to make sure that no leakage appears during the transportation and handling of such products. For such cases, there are no special procedures, the only efficient measure to prevent pollution being the proper and continuous supervision and maintenance of equipment and plants.

VII. METHODS OF CONTROLLING SURFACE-WATER POLLUTION

For the treatment of the noxious materials discharged in water by petrochemical plants, any of the treatment procedures known today may be applied. Since even a brief discussion of these processes would require a large volume, only the classical methods and the most efficient processes will be described here, starting from the most general to the most specific treatment methods.

Mechanical-chemical treatment

All waste waters discharged from the petrochemical plants are subjected to a mechanical and chemical treatment. This type of treatment consists of several distinct steps: separation of large pieces of solid matter, homogenization, neutralization, flocculation and settling.

Large pieces of matter are separated at the inlet to the mechanical-chemical plant by means of metal grills, the material caught by the grill being automatically removed.

Homogenization takes place in rectangular or circular tanks in which the effluents discharged from different processes are stirred and homogenized by means of scraping bridges in preparation for chemical treatment. If material accumulates at the surface or bottom of the homogenization unit, it is removed by the scraping bridge and sent for further treatment.

Neutralization of water at pH 8.5 is carried out with 96 per cent sulphuric acid or sulphuric acid discarded in other processes. Acidic water is neutralized with caustic soda lye or milk of lime when acid recovery is not possible. It is better to use dolomite lime because it produces less slurry and is cheaper than caustic soda or caustic soda lye. The introduction of the neutralizing reagents is controlled on the basis of pH measurements taken with a pH-meter. To carry out a more complete reaction between the neutralizer and neutralizant during the neutralization process, it is desirable to stir the reactants in the medium by air bubbling.

The neutralized water is passed through a unit where a flocculating agent such as ferrous sulphate or aluminium sulphate is added. In certain cases flocculation auxiliaries are also added to increase reagent action. To achieve an intimate mixture between the reagents and increase the reaction yield, air bubbling is carried out in this step also.

Settling takes place in circular or rectangular tanks. The aim is the deposition of the suspensions and of the insoluble salts formed during the neutralization process. When this step is completed, the waste water may be discharged into the water course if it meets all the conditions imposed, or it may be transferred to a higher treatment.

Activated-sludge (biological) treatment:

An activated-sludge treatment is used to reduce the organic content of the effluent by digesting these compounds with the help of the micro-organisms existing in the activated sludge and by oxidizing the products with oxygen. Oxygen is diffused throughout the aeration tank by air bubbling (through stainless steel or plastic tubes) or by means of turbines, the oxidation in this case taking place in the atmosphere at the interface between the air and the water particles. Both processes provide the maximum use of the entire volume of the aeration tank achieving at the same time a homogeneous distribution of the micro-organisms and oxygen in the water. The oxygen is necessary not only for the oxidation process but also to keep the aerobic micro-organisms alive and multiplying. Nutritive agents containing phosphorus and nitrogen may also be added to maintain optimum life conditions.

The biodegradation (biological degradation) time depends on the biodegradability of the products and on their concentration, and it may range from 2 to 10 hours.

The waters together with the sludge from this step are transferred to settling tanks. Some of the sludge that settles is returned to the aeration tank to replace what was used there, and the excess is sent to a sludge-treatment unit. The treated water may be subjected to an additional flocculation and settling process or filtered through a sand filter and discharged into the water course.

The advantages of this method in obtaining high-quality effluents can be obtained on small scale, with a pressure vessel and with a suitable power source.

The disadvantages of this process are high maintenance and operation costs, difficulty in determining the volume, and the sensitivity of the micro-organisms to the acids, or high concentration of noxious materials, difficult recovery of the bacteria cultures are affected by the noxious materials (heavy-metal ions, chlorinated paraffins, etc.) or by shock loadings, the necessity of careful control of the returned sludge volume and concentration, etc.

Sludge treatment

If a high volume of wet sludge having a high content of organic materials is obtained in waste-water treatment, some measures have to be taken for its treatment and storage. The following are the most common methods being applied:

(a) Anaerobic digestion in heated, closed tanks provided with devices to maintain the optimum alkalinity conditions. This system receives the sludge from primary settling (after the mechanical-chemical treatment) and secondary settling (biological treatment). The digested sludge passes to dehydration and storage;

(b) Anaerobic digestion preliminary to the other treatment stages, known as a "pre-digestion" process, which is efficient especially in waste waters having a high biological oxygen demand (BOD) or chemical oxygen demand (COD). Use of this method allows a decrease in the size of the treatment units, especially the settling tank and the biological stages;

(c) Dehydration of the wet crude sludge by physical or thermal destruction in thickeners, in vacuum filters, by spray drying, by centrifugation or by other processes.

When anaerobic digestion is applied, it is necessary to operate with two vessels, the first being used for digestion and the second for the storage and collection of the gas formed. Usually, the first vessel is

fitted with stirring devices to provide an intimate contact between the sludge being digested and the newly introduced bacteria. A digestion of organic material does not take place until optimum temperature and other medium conditions are obtained in the digestion unit.

The heating of the contents is achieved by coils located on the inner surface of the reactor or by vertical coil assemblies suspended from the vessel cover. In this case, the sludge that is stirred by the streams formed by stirring will flow and come into contact with the hot coils. There are also cases when the sludge is heated from the outside of the vessel, using as a heating medium the liquid that has overflowed and is being recycled. The hot water used is obtained in boilers or heat-exchangers located in the manufacturing flow between the two digestion units.

From experience it has been found that a process similar to natural metabolism and requiring a certain temperature and alkalinity takes place in the digestion tanks. The result is the production of methane gas with a low content of hydrogen sulphide that is suitable for combustion. In order to achieve a digestion without odours leading to the production of useful gas, it is necessary to increase the temperature of the digester content to 32-25°C and to use a reagent to adjust the pH to 7.

Since the digestion tanks are large, sometimes 30 m in diameter and 8 m high, a significant quantity of heat is required. This heat can be obtained from the combustion of the gas resulting from the sludge digestion.

After vacuum filtration, centrifugation etc., the sludge is dehydrated and sent to the incinerator (which may be vertical with superposed hearths, tubular like a cement oven, in a fluid bed with inner ceramics or sand, or some other construction) fueled by methane, residual gas or heavy oil.

Dialysis and electro-dialysis

Dialysis is the separation of substances in solution by means of their unequal diffusion through a semipermeable membrane. Electro-dialysis, or dialysis in the presence of an electrical field, is particularly well suited for the separation of the ionic components from a solution. The

water is drawn from the more concentrated, the membrane with the pores between, the water is forced from the dilute to flow. One membrane permits only water to pass through and not salt ions; the other, with its pores smaller, permits only ions. As in an electrolytic cell, the anode attracts the positive ions and the cathode, negative ions. The cell is built so that the water that is attracted to and permeating the anode and cathode membrane is returned to the solution passing between the membranes. Thus, the solution subjected to electrolysis becomes highly purified.

This process is quite efficient but very costly because of its energy consumption of 10-15 kWh/m³. Moreover, the process requires an effective removal of the salt ions from a highly concentrated solution that collects in the membrane. For this reason, membranes of different kinds of membrane must be used, depending on the kinds of salt to be removed.

Reverse Osmosis

Reverse osmosis is a kind of filtration by membrane sometimes called "superfiltration" or "hyperfiltration". It differs from dialysis in that high pressure is used to force the liquid through a semi-permeable membrane, leaving behind the solid suspensions or the dissolved salts. In normal osmosis, the flow is the other way: pure water passes from the less concentrated solution through the membrane into the more concentrated solution. When pressure is applied to the more concentrated solution, the water passes in the reverse direction; thus the name, reverse osmosis.

Originally, this method was used for desalting sea water. Now, it is used in industry, to remove sulphuric acid and hydrogen chloride from their aqueous solutions. The process is relatively expensive but is the best solution for treating water on which severe restrictions apply or that should have a high purity for recycling.

In some plants, the system removes, in one step, solid suspensions, bacteria and viruses from the effluent. After this process, the BOD of the effluent is 5-10 mg/l. The results are comparable to those of biological treatment, but the cost of the plant using a filtering membrane is lower than the cost of a classical treatment plant, where the purpose is to obtain high purity water.

The membranes used are cellulose acetate strips, which are not easily blocked by suspensions and allow flow rates of 400-1, 000/40/70 centimetre surface at pressures below 1.5 kg/cm². The plant is not too compact and occupies little space; its construction is installed inside the building. Each cubic meter of plant provides a filter area of 150-200 m².

The process described has given satisfactory results in the treatment of waste water from chemical and pharmaceutical industries and has also been successfully applied in the treatment of water containing large amounts of organic impurities.

Chlorine, hypochlorite and ozone oxidation

In some cases, the process waste waters may still be heavily loaded with organic matter even though they have been subjected to mechanical-chemical and biological treatment, so that they do not meet the standards for discharge into a public water course. A finishing step using strong oxidants such as chlorine, hypochlorite or ozone is carried out. This oxidation process is also applied to water containing cyanides and phenol. Usually ozone is used in this case because it oxidizes cyanide ion into cyanate ion, which is about 1,000 times less toxic and hydrolyses to CO₂ and NH₃. In the case of phenol, ozone is used in the finishing step in order to avoid the chlorophenol that would be formed in the oxidation by chlorine or hypochlorite. It is known that chlorophenols give an unpleasant taste to drinking water when present in concentrations as low as 0.015 ppm.

In some countries, ozone is used in treating the waste water from refineries and coke ovens (such water is similar to the water coming from petrochemical plants). Phenols, as well as mercaptans and other sulphur compounds, disappear after being in contact with ozone for only 3 min. After 15 min of contact, an organic material content of 30-50 mg/l is reduced to 10 mg/l. The ozone consumption in this case is 1.5-2.7 mg/mg contaminant.

In making the ozone if oxygen is used as raw material instead of air, twice as much ozone is produced for the same consumption. Either raw

material must be dried at -60°C and fed into the ozonizer at a temperature below 11°C . It is recommended that 99.5 per cent pure oxygen be used.

The most practical ozone/contaminant ratios are the following: 1.5-2.0 for phenol and 1.0 for cyanide ion. In treating refinery waste water, the goal in the tertiary treatment is to obtain a final effluent with maximum phenol content of 3 ppm.

The over-all energy consumption of the process is about 16 kWh per kg of ozone.

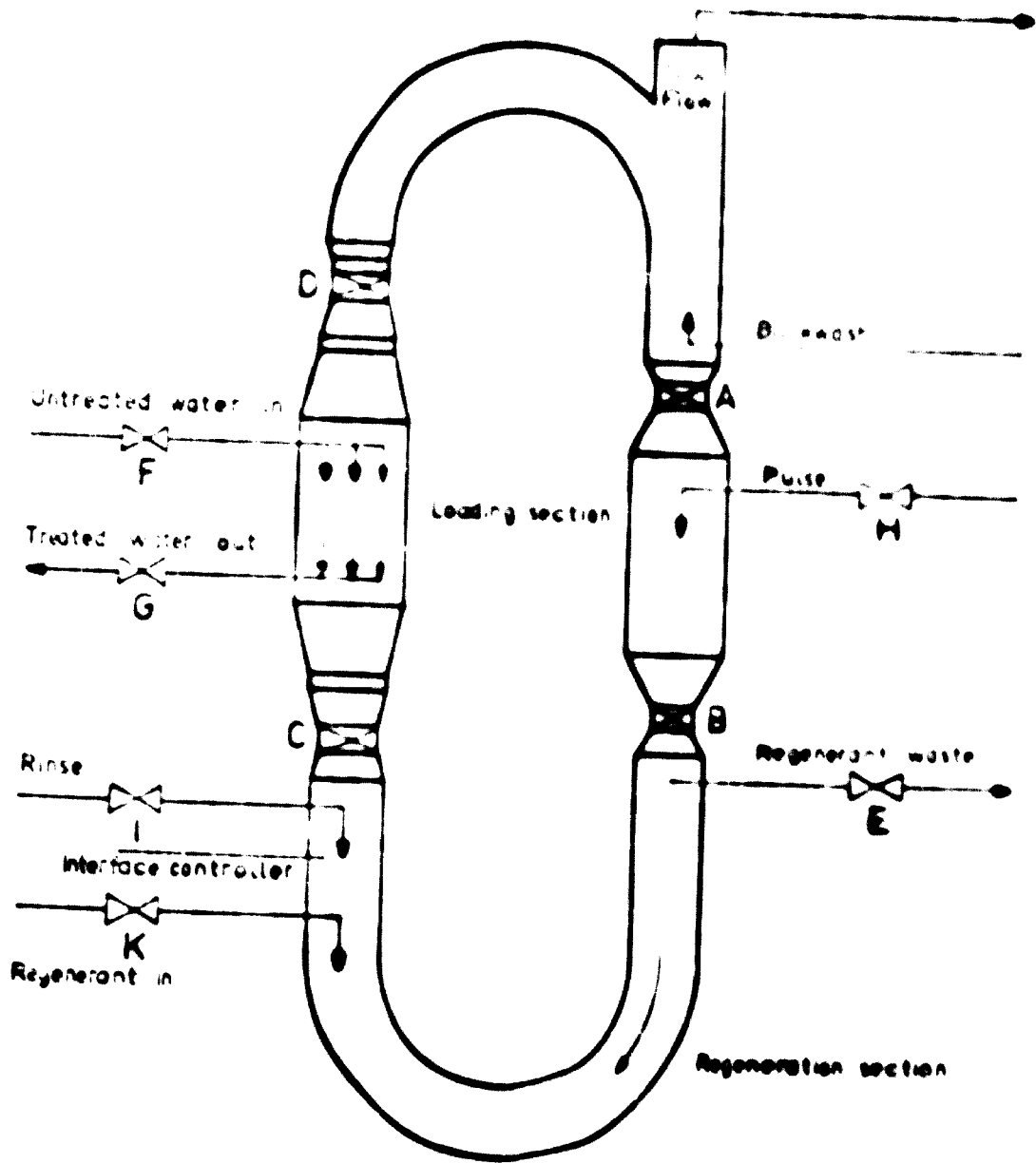
Treatment of water containing nitric and ammonium ions

Formerly, it was believed sufficient to neutralize water containing nitrate and ammonium ions with lime or acids, but experience proved the process to be inefficient because of hydrolysis. Several other treatment and destruction processes have been developed, among which are stripping and ion exchange.

Stripping of waste water containing these contaminants is carried out with low-pressure steam in ordinary towers after a preliminary alkalization of the effluent. Ammonia is extracted at the top and returned to the process or utilized as a fertilizer (ammonia water), and the acidic water from the still is neutralized.

Ion-exchange adsorption was once an expensive batch process. Now, a continuous process is available, using a reactor similar to that shown in Figure 1. The process has four distinct stages. In the first stage, the ion-exchange resin is put in contact with the ion that must be adsorbed. When the resin is saturated, admission of effluent is stopped and the resin is passed to the following section where it is washed. In the third stage, regeneration of the resin is carried out with the recovery of the adsorbed ion in the form of ammonium nitrate, and after this the resin is transferred to the fourth stage for washing loosening before return to the first stage of the process. Any resin lost in the process is replaced in the last stage.

FIG. 1. Schematic diagram of a two-stage ion exchange process.



Valve positions during cycles:

Load cycles: valves A, E, F, G, I, K open; valves B, C, D, H closed.

Rinse cycles: valves B, C, D, H open; valves A, E, F, G, I, K closed.

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Mineralization

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Final

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This process is actually an alternative of the activated sludge method:
It depends entirely upon mechanical aeration and stirring of the effluent
to maintain a turbulent medium in the ditch and to ensure contact between the waste
liquid and the aerated oxygen. This is achieved by an airer spraying
the water on the ditch surface. By this method it is possible to
accomplish complete aeration and to lengthen the aeration period for a
long-term oxidation (usually longer than 24 h). Sufficient oxygen is
provided for the stabilization or destruction of the phenolic product.
The volume of an oxidation ditch should be 5 m^3 for 1 kg BOD per day. A
liquid layer thickness of 0.90-1.5 m is recommended. The ditch may be
in the form of a ring or a single or double ellipse but must form a closed
cycle. The median strip must be wide enough so that the ends of the ditch
should not be too acute. To handle 40 kg BOD per day, the median strip
should be 1 m wide.

Besides the ditch, there is also a mechanical-aerated water treatment at the inlet and a settling of the water discharged from the ditch. Part of the sludge from the settling tank is returned to the ditch and the excess is deposited in dry pits and treated by one of the processes described above under "sludge treatment".

Phenolic waters may also be incinerated. This method may be applied to effluents with very low flow rates and very high phenolic concentrations or to phenolic waters. The process is expensive because of its high fuel consumption.

The incinerators are vertical kilns that have four, six or eight burners at the bottom arranged on the furnace circumference and a central burner. The effluent to be incinerated is sprayed through two to four nozzles onto the flames of the burners. The burning regime is controlled by thermocouples and may be automatically adjusted.

Cyanides

As was previously pointed out, ozone is preferred to chlorine or chlorine compounds as a reagent for destroying cyanides. The cost of a plant using ozone is higher, but the operation is cheaper. In literature one finds mention of a consumption of 2 kg ozone per 1 kg cyanide, at a cost of about \$1, as compared to other plants where 8 kg chlorine are used for 1 kg cyanide at a cost of about \$1.70.^{12/} Hypochlorite is not suitable for high cyanide concentrations and volumes because it is not much cheaper and requires complex equipment for operation and handling. Also, hypochlorite decomposes easily and becomes useless. The cost of ozonization plants decreases considerably if they are operated continuously or are not overdimensioned.

As with phenolic waters, when the effluent containing cyanides has a low flow rate and high concentrations, incineration may be adopted as a treatment method.

^{12/} Edmund B. Besselievre, The treatment of industrial wastes (New York, McGraw-Hill, 1970).

Hydrocarbon

Removal of hydrocarbons from the effluent is achieved with high yields (90-95%) by a parallel-plate separator. For this purpose, the water is treated with a flocculation tank.

In place of a flocculation tank may appear an "absorber" (high-volume surger), provided that it is installed ahead of the homogenization tank to absorb the hydrocarbons. After homogenization, the waste water is subjected to a parallel-plate separator. The most common types of separator are: (1) the type standard used by the American Petroleum Institute (API), the parallel-plate separator (PPI) and the corrugated-plate separator (CPI).

The API tank is of a size calculated to provide a slow velocity of liquid passing through the tank and a uniform distribution of the liquid over its surface. The construction enables the oil and hydrocarbon particles to form a thin layer on the surface while the heavy materials fall to the bottom. A built-in skimmer shoves the hydrocarbons into a float and the heavy materials into a tank. The most important factors in the yield of the API tank are circulation velocity, time required for the oil to separate from the water, suspension settling rate, water temperature, density and diameter of the oil drops, quantity and composition of the suspended matter, presence of sulphur and detergents, pH and geometry of the tank.

The PPI tank is a conventional tank divided into longitudinal channels by some parallel plates. In this way, the separation capacity is increased because the maximum height that an oil drop must rise before reaching the surface is reduced. This separator also has the advantage that it diminishes turbulence and favours gravitational separation. The process is less expensive than the one described previously.

The CPI tank is like the PPI tank, except that its plates are corrugated rather than flat. This fact leads to two interesting phenomena: sliding of the sediments downwards in the circulation sense and regrouping of the oil particles upwards in the counter current. This device provides a slight improvement by increasing the hydrodynamic quality and creating counter currents between the water to be treated and the hydrocarbons, thus favouring coalescence.

This new element, which used in the 1950s, is a process of water hydrocarbon concentrations in the range 1-2 g/l, which are reduced to a hundredth of a liter per liter of hydrocarbon. This is achieved by a process. Therefore, after this treatment, if the water contains a large amount of products, it may be discharged directly into water bodies. 13

Other mechanical-physical processes for removing hydrocarbons are settling, preceded by active flocculation and filtration.

Settling, preceded by active flocculation, requires the consumption of a large part of the suspended matter and hydrocarbons. This treatment is doubled by a removal treatment of suspended matter being dealt with. Demulsification by chemical precipitation when the sulphide or mercaptan content is below 100 mg/l is carried out efficiently by utilizing chemical coagulant reagents such as iron or aluminum salts. At the same time, if hydrocarbons are present, it is necessary to increase their weight by adding lime and diminishing the settling rate. In this case, floating is desirable.

Air floating and electrofloating is carried out only after the addition of an agent to the liquid to be flocculated, so that the microbubbles that collect around the hydrocarbons and suspended matter may carry them to the water surface to be removed by a scraper. The microbubbles are produced either by the release of pressure on an effluent saturated with air at a pressure of 4-5 bars (air floating) or by the release of oxygen and hydrogen bubbles from the surfaces of immersed electrodes (electrofloating). The flow rate of liquid in the floats is higher than that used in settling tanks and varies from 3 to 6 m/h depending on the nature of the water, the degree of treatment and the concentration of the floating agent. It is important that the distribution of the microbubbles be as uniform as possible and the hydraulic concentration be controlled. In air-floating, the lime consumption varies between 20 and 50 l/m³ water, the flow of process water representing 20-40 per cent of the treated water flow.

13/ "Techniques pétrolières et pétrochimiques modernes, leurs effets sur la production d'eaux résiduelles" - conference presented by the French Government at a seminar in Warsaw, September 1973.

(Recycled water can be used for process water). A 70 per cent decrease of the organic suspended matter and a 35 per cent decrease of the TOC are to be obtained.

In the case of a secondary effluent, the hydrocarbons that remain after the primary treatment may be destroyed in the biological treatment or by a further treatment with oxidizing agents and filtering layer.

Effect of toxic substances

The toxic substances mentioned in chapter IV are preliminarily tested in the laboratory on test organisms, on which the possibility of eliminating them by biological treatment depends. At the same time the toxicity of the products of the last step is checked.

Irrespective of the test results, waters containing organic materials are subjected to a mechanical-chemical treatment. If they are biodegradable, the effluent of the biological step, otherwise, they are subjected to reverse osmosis, electrodialysis or oxidation. If the materials are not biodegradable, a filtered separation is attempted and a low flow rate is maintained together with a high concentration to prepare the effluent for incineration.

Heavy-metal ions

As was mentioned before, zinc, chromium, copper, manganese, nickel, cobalt and vanadium ions are strong inhibitors of the biological treatment process. Therefore, the treatment of water containing these elements is carried out quite separately from that of other water in chemical plants. This treatment process is relatively simple and is based on the precipitation of these ions by adding a dilute sodium hydroxide solution (10 per cent) simultaneously with a flocculating agent (ferrous sulphate) and stirring the solution vigorously. After the reactions take place, the hydroxides are separated, removed by filtration through a press filter and deposited in the dump. The water is discharged into the storm sewer. It must be pointed out that not all the metal hydroxides precipitate at pH 6-7; chromium, copper and manganese precipitate at pH 7-8 and nickel at pH 7.5-8.

Mercury salts

The treatment of water containing mercury salts is also carried out separately. In this case, the mercury is precipitated with calcium hydroxide and carbon disulphide; after filtration, the water is discharged separately into the storm sewer. The carbon disulphide is regenerated by distillation, and the mercury is recovered and returned to the process. The wastes are put into sealed containers and deposited in specially arranged places (surge tanks reinforced with concrete).

Carbon black

The removal of carbon black from waste water is effected by mechanical-physical processes. Dry carbon black (high specific surface) is retained in the settling tanks, where it rises to the surface with the light hydrocarbons and other light suspensions, collected and incinerated. Wettable carbon black is found on the bottom of the settling tank and is discharged with the sludge to the sludge thickeners and fermenters.

When the amount of carbon black is quite large (e.g., in partial oxidation methane plants), the water is separately treated, the carbon black being separated according to the above-mentioned principles and incinerated after it has been dehydrated by filtration and compression. If the settling process does not give satisfactory results because of the presence of petroleum products, the air floating process is recommended.

VIII. METHODS OF CONTROLLING SOIL AND GROUND-WATER POLLUTION

To many people, the term "waste deposit" still connotes large areas chaotically covered with various domestic and industrial wastes. In fact specialists in this subject have worked out a series of principles for the salubrious and organized deposition of wastes.

To define the deposition area, data on the following items are necessary:

- Geological structure of the land
- Characteristics of the ground-water layers
- Ultimate disposition of the waste

For the first two items, it is necessary to carry out some drillings and detailed analyses. The former requirement for a low soil permeability in the deposit area is no longer so important because of modern techniques (insulation with reinforced sheets or bitumen, draining systems etc.).

The organized deposit is accomplished with industrial equipment within a well defined perimeter and under continuous supervision, in order to avoid the scattering of the waste. The soil water is continuously examined and analyzed to detect any exfiltrations from the deposit. The organized deposit may also include provision for reclamation of the land after the deposit is broken up. Depending on the structure of the soil, the area may be used for agriculture or recreation.

Construction of waste pits

Waste deposits are usually located on non-fertile areas around the industrial chemical works and, if possible, where impermeable loess or marl layers exist at the surface or slightly below. In this case, the bottom of the storage pit need not be impermeabilized. The marginal dikes are made impermeable shields of impermeable gel. On slopes, impermeabilization is effected with stabilized clay protected with sand.

In case the latter is permeable, or the top, usually lower, is, asphalt layers may with suitable thickness be used to prevent them, being reinforced with glass fibre, or used to impregnate the deposit. For less severe conditions, only top polyethylene sheet need be used in tests with the waste, and they will be adequate. For extremely toxic wastes, concrete-reinforced pits are recommended.

Discharging of wastes at the disposal site by overhead cranes with grab buckets or conveyor belts. In the case of large deposits, access roads must be built.

Determination of groundwater movement

It is necessary to determine the degree of ground-water pollution by the noxious materials that may eventually exfiltrate through the unsuitably storage pit systems into the ground-water system. A radiometric method has been developed^{14/} for the establishment of the direction and rate of flow of water and porosity. Its principal advantage is that it does not disturb the natural conditions of the water flow at the given place and during the tests. One starts from the hypothesis that a water volume marked with radio-nuclide will have the same dynamic parameters as a non-marked water volume. For determining the velocity vector, it is necessary and sufficient to determine at any moment the position of the centre of gravity of the radio-nuclide cloud.

The radioactive Br^{82} solution is introduced in the area by a vibratory pile driver. The tracing of the movement of the radio-nuclide cloud through the investigated porous medium is done by detecting sounders on the circumference of a circle centred on the injection point and with a radius dependent on the supposed water velocity. The detection of the centre of gravity of the radioactive tracer is done by sliding a Geiger-Müller radiation counter inside the detecting sounder and recording its counting rate at different distances.

It is possible to determine the flow direction by tracing the centre of gravity of the radioactive tracer. The radio-nuclide cloud carried off

^{14/} "Radiometric method for the determination of velocity, flowing direction and course of the ground waters", Romania Patent No. 47.215-1965.

ly the sand-water will come close to the detecting scanner or location in the direction of its movement. The counting rate in this scanner will be a maximum when the centre of gravity passes through that point. The horizontal component of the flow direction is given by the straight line joining the launching point with the location point of the detector scanner registering the maximum activity. The vertical component, being the depth of the centre of gravity, is obtained from an analysis of the counting rate variation with depth in the same scanner. Reproduction of the whole experiment may be extended to greater distances by extending the horizontal distance and replacing it further along the projected direction of flow.

The flow rate is determined by direct division of the distance covered between the moment of launching and the moment of detection of the maximum radioactivity by the time interval between those moments. However, there is a difference between the tracer speed and the actual water flow rate, caused by the interaction phenomena between the radio-nuclide and the porous medium that was penetrated.

Analysing these phenomena (adsorption and absorption, ionic exchange, diffusion etc.), many researchers have concluded that the difference between the tracer and water speeds, if not adjusted for, may lead to relative errors as high as 50 per cent, especially for low water flow rates.

Applying a differential equation for the rate of change of tracer concentration in a porous medium derived by Inoue and Kaufmann^{15/} to the case of cation exchange, Vermeulen and Hierster^{16/} arrived at this relation:

^{15/} Y. Inoue and W. Kaufmann, Health Physics (1953).

^{16/} T. Vermeulen and N.K. Hierster, Industrial and Engineering Chemistry (1952), p. 636.

$$\frac{v^a}{v^t} = \frac{1}{1 + \frac{qP}{C^0}}$$

where v^a = water speed

v^t = tracer speed

q = ion exchange capacity of the medium

p = density of the medium

C^0 = initial concentration of the tracer

P = porosity of the medium

The method and calculations have received practical application in Romania, where it has been found to give very accurate results.

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Once the limit values for noise are decided, it is necessary to establish the permissible exposure time of the individual, considering that at 90 db the usual work begins to be disturbed and at 70 db it is very difficult to continue such work.

Dependent on noise level and exposure time, the sound pressure level at all levels of the population is relatively high. The noise pressure level is high at night, a decrease of the noise level is observed at night and the sound pressure level is high at night.

To simplify the calculation of noise pollution in a plant, one may assume that all the noise sources in the plant are treated together at a noise point called the plant polluting centre. This approximation is valid for average plants with blocks of houses below 50 m in size and the observation point located 50 m away.

If L_{pA} is the maximum permissible sound pressure level (established by standards), at point A located at a distance r_A from the plant polluting centre, the maximum pressure that should be produced at the plant polluting centre LMA is

$$LMA = L_{pA} + 10 \log (\pi r_A^2)$$

This relation is based on the assumption of propagation by spherical waves in a calm and isotropic atmosphere. It does not take atmospheric absorption into account, which is proportional to distance and depends on frequency. There is also a wind effect that modifies the shape of the wave surface. This phenomenon leads to an increase in loudness at observation points situated in the wind direction and a decrease at observation points in the opposite direction. To correct for this in cases of an almost constant prevailing wind, albeit imperfectly, one may introduce a correction independent of frequency of ± 5 dB at 400 m and ± 10 dB at 1000 m. For plants in construction, the screening effect of the buildings should also be considered.

When there are several sources of noise, it is also advisable to classify the sources principal frequencies of the sounds they produce: low frequency (≤ 350 Hz), average frequency ($350 \leq f \leq 1,700$ Hz) and high frequency ($\geq 1,700$ Hz).

X. TWO SPECIAL EXAMPLES OF AIR-POLLUTION CONTROL IN FIBROUS CELLULOSE

The following two examples are given as two special examples of pollution control, and do not in any way, constitute a range of measures which should be employed for the discharge of, say, a mill.

Example 1: The production of acetylene

Background

The production of acetylene with the complex process the acetylene obtained by the pyrolysis of the production of vinyl acetate and polyvinyl acetate (PVA).

The main source of pollutants released ammonia and ammonium salts, nitric acid, and products containing arsenic with an effluent flow rate of 250 m³/h.

The methane pyrolysis and PVA plants release organic materials at the rate of 0.8 t/h, mainly naphthalene, anthracene, ammonia and hydrocarbons. The effluent has a flow rate of about 1,000 m³/h and a BOD of 1,500 kg/h.

The conditions imposed on the effluents discharged into the river are as follows:

BOD	max. 10 mg/l
COD	max. 150 mg/l
Ammonia	max. 0.3 mg/l
Hydrocarbons	max. 0.1 mg/l
pH 6.5 - 9.0	

Ammonia and Nitrate Treatment

Ion exchange treatment

The effluents that contain ammonia and nitrate (condensate water - NH_3 and NO_3^-) are treated by ion exchange resin to remove nitrate and chloride ion, which are then later removed from resin with ammonia solution. The remaining ammonia in the effluent solution is passed to a third step where a stream of air is blown through the solution (to about 10 percent), the ammonia content is then transferred to the ammonium nitrate plant and returned to the normal process operations.

The acid or alkaline waste water that come out from the ion exchange resin are sent to an auxiliary neutralization plant and then discharged into the river.

Stripping treatment

Simultaneously with the first process, the effluents with low ammonia content and high flow rate that appear continuously are treated continuously in the fertilizer plants are stripped alkaline with sodium hydroxide in excess to completely release the ammonia. This process takes place at 100°C with steam in the distillation tower. The ammonia and water vapours are condensed in the form of ammonia water having a concentration of 15-20 per cent ammonia. This ammonia water is reused within the same complex in other manufacturing processes and in the biological treatment.

Arsenic

From the carbon dioxide washing towers in the ammonia plant, the following waters with arsenic content are obtained: continuous condensate, rain and washwater from the platforms, waste water resulting from accidental

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... .. collecting tank from which
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... .. may be sent
... .. or used in the transportation
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A large amount of carbon black and coke is separated from the cooling water in the lower part of the pyrolysis furnaces and sent to a rotary filtering unit. The filtrate is discharged into the storm sewer and the moist carbon black and coke residue on the filter is incinerated in rotary ovens.

Ammonia and higher acetylenes from the methane pyrolysis

Since the effluent emerging from the acetylene plant contains both ammonia and the higher homologues of acetylene, such waste waters are incinerated in a fluid-bed furnace with two combustion sections. For the incineration two furnaces are provided and they operate alternately.

In the lower section, which is shaped like a truncated cone and has a finished consisting of refractory lumps or gravel, the water is evaporated and the residue burned, incompletely or completely depending on the quantity of primary air introduced. The temperature is in the range 800-1100° C.

In the upper section, complete burning takes place in the presence of secondary air introduced tangential through two holes, 180° apart. The temperature is about 800° C.

The methane-gas burner that initially produces a temperature of at least 800° C and maintains it during the waste incineration is placed in the lower part of the evaporation-burning area.

Pollutants from polymer production

Pre-treatment

The pre-treatment of the waste water from monomer and polyvinyl acetate production consists of a polymer suspension separation and a pH adjustment of the discharged water with calcium hydroxide. The polymer suspension, settled in a local settling tank, is extracted and deposited separately in an appropriate storage pit. The water is then given a mechanical-chemical treatment.

Mechanical-chemical treatment

The water first passes over a screen where the coarse material is separated out and mechanically removed. Further on, the water passes into a sand-clearing basin. The separated sand is discharged in batches by a mammoth pump. The water is treated with calcium hydroxide and directed to the primary settling tanks. These are fitted with hydro-elevators for recycling the sludge and scraping bridges removing it. They also have collecting systems for the hydrocarbons that happen to appear after local treatment and separate out in this stage. The lime-milk suspension is fed by pumps controlled by a pH-meter. The sludge that was separated in the settling tank is removed by gravity to the pumping station and then to the ash dump.

The resulting effluent, which is a BOD-free liquid, is collected in a return system and the pH is adjusted with sulphuric acid under agitation by air bubbling. The precipitation of the calcium sulphate formed takes place in a settling tank with a scraper bridge. The sludge is sent to the sludge dump.

The treatment yield of these two steps is 90 per cent of the organic content (BOD). For special situations (acidic water), either step may be substituted by the other.

Homogenisation, the next step, is provided for both steady and irregular flows (e.g. rain). It is accomplished by means of two tanks operating alternately or in parallel. The tanks have aeration turbines mounted on floats in order to follow the level variation. The homogenised water is passed to a second precipitation tank and then to the biological treatment station.

Biological treatment

Nutrients containing phosphorus and nitrogen are added and the water and activated-sludge mixture is aerated by 40 turbines provided with suction tubes. Of the total volume of the tanks (41,500 m³), 75 per cent is for the biological treatment and 25 per cent for activated-sludge regeneration. With an over-all oxygen consumption rate of 2,200 kg/h, a BOD of 0.605 kg/m³/day can be handled.

The phosphorus and nitrogen required for the nutrition of the micro-organisms are given in the form of sodium phosphate and ammonia water, respectively. The proportioning of the nutrients is carried out automatically depending on loading and flow rates. The yield of the biological treatment is 90 per cent of the input BOD.

Secondary settling

The water and activated sludge mixture is gravitationally guided to the secondary settling, accomplished in circular settling tanks equipped with scraper bridges. The separated water is passed to the third step - a mechanical-chemical treatment.

The separated activated sludge is collected by a special tube system mounted on the scraping bridge and is directed to a central metallic tank from which it is recycled to the aeration tank by means of vertical pipes. The sludge excess is removed by gravity to the pumping station and deposited in the ash dump.

Tertiary mechanical-chemical treatment

In order to assure a high degree of treatment, a final flocculation-settling step is provided. The equipment consists of the necessary reaction vessels and two tertiary settling tanks. The treated water is drained through a flow-metered channel to the treated-water collector, and the chemical sludge formed is discharged gravitationally to the pumps and further on to the ash dump.

Corrosion problems

Corrosion problems appear in these plants, especially at the ammonium-ion recovery unit, where special acid-resisting steel must be used. In the monomer and polymer plants, the effluent settling tanks and catchment basins are lined with acid-resisting bricks and the fittings, joints, pipes and pumps are manufactured of special acid-resisting steel. In the mechanical-chemical and biological treatment, the tanks are protected by an epoxy-resin coating.

The influent conduits for the industrial waste water are made of glass-fibre reinforced polyester, and the reagent preparation, proportioning, feeding and transportation equipment is manufactured out of special steel.

Process control

The entire waste water treatment process is controlled from a central control room where all the data is recorded by instruments. Depending on this data, the plant may be operated automatically or not by remote control.

Second example: A petrochemical complex consisting of monomer production and processing plants and an oil refinery

The second example analysed is a petrochemical platform consisting of a refinery, pyrolysis plant, plants for the separation of monomers and production of polymers, polyethylene, polypropylene, polystyrene, complementary plants for the manufacturing of other products (acrylonitrile, ethylene oxide, glycols), and a carbon-black plant.

Effluents:

The main contaminants that may be found in the waste water are the following:

Inorganics: sodium hydroxide, sodium sulphate, sodium carbonate, calcium chloride, sulphuric acid, titanium dioxide

Organics: benzene, ethylbenzene, distyrene, ethylene glycol, ethyl alcohol, methanol, N-methylpyrrolidone, cyanogen chloride, hydrocyanic acid, gluconates, dichlorhydrin, methanamine, acrylonitrile, oils, polymer suspensions

The average hourly discharge of the effluent is 1400 m³ and the BOD load is 13,000 kg/day.

The limit values imposed on discharges into the river are the following:

BOD	10 mg/l
COD	150 mg/l
sulphides and hydrogen sulphide	0
phenol	0.02 mg/l
mercaptan	0.5 mg/l
oils	0.3 mg/l
pH	6.5-8.5

On the basis of these data, the following sequence of treatments was adopted for the treatment of the waste water discharged from this petrochemical complex:

- Primary (mechanical-chemical) treatment
- Secondary (biological) treatment, in two steps
- Tertiary finishing treatment (physical-chemical)
- Sludge treatment

The treatments comprise the following operations:

Primary treatment

- Homogenisation
- Neutralization
- Oil separation
- Flocculation and primary settling

Secondary treatment

- Aeration 1
- Sludge regeneration
- Secondary settling 1
- Aeration 2
- Secondary settling 2

Tertiary treatment

- Secondary flocculation - tertiary settling
- Filtration through sand

Sludge treatment

- Thickening
- Filtration
- Incineration

Description of the treatment operations (see Figure II)

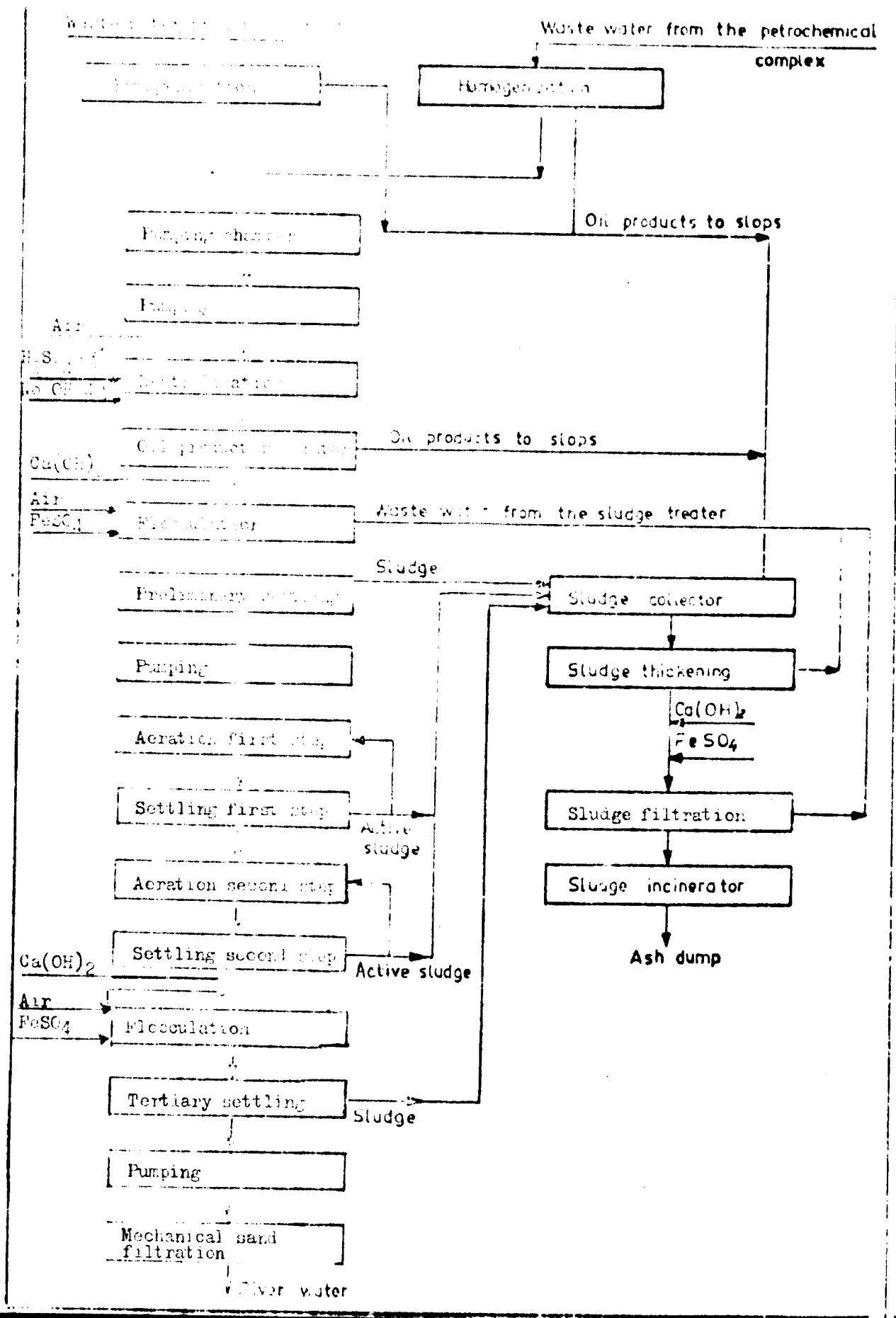
Primary treatment

Homogenization of the waste water from the refinery and from the petrochemical complex takes place separately. This first stage also includes a primary separation of the oil products, which are directed towards the slop separators. The homogenization tanks are circular in shape, have a scraping bridge and are calculated for a homogenization time of 4-4.5 h. The homogenized effluents are discharged through two pipes to a single pumping chamber. Their flow rate is automatically maintained constant.

Since in this case the two homogenization tanks discharge at a lower level than the neutralization unit, it is necessary to use pumps - in this case helical pumps to convey their discharges to neutralization unit. If the ground permits, situations like that should be avoided to save the expense of pumping.

Neutralization takes place in a rectangular tank by adding sufficient 96 per cent sulphuric acid or 40 per cent sodium hydroxide to make the water pH at the outlet 6.5-7.0. The tank contents are stirred by air blasting. Average neutralization time is 10 minutes.

PLANT LAYOUT OF THE WASTEWATER TREATMENT PLANT OF THE PETROCHEMICAL COMPLEX
 (WATER TREATMENT PLANT OF THE PETROCHEMICAL COMPLEX)



The neutralization reagents are automatically delivered from their storage tanks by proportioning pumps according to the indications given by the pH-meter.

The neutralized water then passes into a circular oil product separator, 2.5 m in diameter, made of reinforced concrete and provided with a scraping bridge to collect the oils into a hole on the periphery, from which they are gravitationally discharged to the final oil product separator - slope.

The separated water flows through several slits made in the separator wall and into the flocculator. At the discharge points, lime milk is automatically fed into the water by proportioning pumps to bring the pH up to 10.5; the residence time is about 50 minutes.

Flocculation and primary settling is achieved in a circular tank, concentric with the oil product separator and divided into sections, where the process operations described below take place. The inner diameter is 2.5 m, the outer diameter, 30 m.

In the first circular section, the orthokinetic step of the flocculation is effected under fast stirring with an amount of ferrous sulphate automatically determined by the flow rate. Then, the water passes into the second concentric section where the perikinetic step under a slow stirring takes place and the flocs are formed. The flocculation time is 10 minutes.

After flocculation the waters pass through a flattening and quieting device into the primary settling tank, which is concentric to the petroleum product separator and flocculator, and come out through a collecting drain into a surge tank. Inside diameter is 30 m, outside diameter 45 m. An extension of the scraping bridge of the petroleum product separator works simultaneously in the primary settling tank, collecting the sludge into a central hole. The settling time is 100-110 minutes.

The collected sludge is passed to the sludge thickener by vertical pumps. In this primary treatment, a decrease of 4 per cent in the initial BOD and 20 per cent of the initial COD is obtained.

Secondary treatment

In the secondary (biological) treatment, the organic load is decomposed by the action of micro-organisms. Since the water is highly polluted, the biological treatment consists of two aeration-settling steps.

Aeration 1. The water is pumped by vertical pumps from the primary settling tank into the suction system of the aeration tank, a rectangular basin divided into six compartments, into which the water is distributed by a distribution system. The aeration is accomplished by the action of a mechanical aerator.

The nutrients, which are added into the pump suction tank, consist of phosphorus in the form of trisodium phosphate and nitrogen in the form of ammonia, to a concentration of 1 per cent phosphorus and 5 per cent nitrogen depending on the eliminated BOD. This step has 70 per cent yield for a total suspended solids of about $1,400 \text{ mg/l}$, total volume about $5,800 \text{ m}^3$, sludge concentration 4 kg/m^3 , specific oxygenation $2 \text{ kg O}_2/\text{kWh}$, and overall consumption rate of oxygen about $1,000 \text{ kg/day}$. The aerated waters come out through a venturi-level device and fall freely into the first set of secondary settling tanks.

Sludge from the secondary settling tank is recycled into the aeration tank by a submergible pump and regenerated by aeration with the turbines, the sludge being distributed among the aeration compartments by discharge chutes.

Secondary settling 1. The biologically treated water with about 4 kg/m^3 activated sludge is usually introduced into two secondary settling tanks, 30 m in diameter, fitted with scraping bridges and fast sludge drains. The settling time is about 2 h and 15 minutes. The water is drained through a peripheral refuse spout into a surge tank, from which they are guided toward the second aeration tank.

The sludge is collected by the scraping bridge into a central tank where it falls freely into a surge and is sent as recycled sludge back to aeration 1 or as excess sludge to the thickener.

Aeration 2. The water from secondary settling tank is distributed to the eight aeration tanks of aeration 2 through discharge channels.

As in aeration 1, aeration is by turbines, and sludge is recycled by helical pumps from secondary settling tank. After an aeration time of 4 h, the yield is about 90 per cent. The over-all surface of the compartments is $1,200 \text{ m}^2$, total volume $7,000 \text{ m}^3$, sludge concentration 1 kg/m^3 , specific oxygenation $1 \text{ kg O}_2/\text{kWh}$, and over-all consumption rate of oxygen $1,000 \text{ kg/day}$.

The aerated waters pass through a constant-level device and fall freely into secondary settling tank 1.

Secondary settling 2. The biologically treated water, containing 3 kg/m^3 activated sludge, axially enters a circular tank, 45 m in diameter, for secondary settling 2, also fitted with a scraping bridge and a fast sludge train. The settling time is 3 h and 30 minutes.

From a peripheral refuse spout, the water is guided through a surge tank towards the tertiary settling tank. Within the surge tank, the water is alkalinized with lime milk to pH 9.

The sludge from settling tank 2 is gravitationally drained into a surge tank, from which it is distributed as recycled sludge to aeration 2 by pumps or gravity and as sludge in excess to the sludge thickener.

Tertiary treatment

Secondary flocculation - tertiary settling. The alkaline water from secondary settling 2 enters a surge tank next to the tertiary settling tank, where the orthokinetic stage of flocculation with ferrous sulphate under fast air stirring takes place. The water, mixed with the reagent, axially enters the circular tertiary settling tank, 12 m in diameter, where the perikinetic stage of flocculation takes place (duration 12 minutes). The air required for vigorous water stirring and oxidation of divalent iron to trivalent iron is provided by an air blast mounted on the scraping bridge.

From this section, the waters pass through a flattening and quieting device in the settling tank proper, 35 m in diameter, where the sludge is scraped into a concentric hole and drained hydrostatically towards the sludge thickener. The settling time is 90 minutes.

... .. into
... ..

... .. front
... .. river.

... .. the
... .. the river. The
... .. the filtration rate is about

... .. the working of the
... .. after the
... .. three air blaste
... .. platform.

Sludge treatment

... .. about 2,700 t/day. It
... .. primary, secondary (aeration) and tertiary
... .. three distinct operations involved in sludge treat-
... .. and incineration.

Thickening. The sludge enters the thickener by gravity after it has
been collected in a surge tank from the above-mentioned sources.

After a long settling of about 6 h in a parallelepiped thickener having
a volume of about 1,000 m³, the sludge reaches 4 per cent concentration and
is collected in a central hole by means of a scraping bridge, the water
being returned to the primary step of the treatment cycle.

Filtration. Further on, the sludge is pumped to the filtration unit,
where, after addition of 15 parts of lime milk and 20 parts of ferrous
sulphate to 100 parts dry sludge, it is filtered through six rotary filters
having a total surface of 240 m², and attains a concentration of 20 per cent.
The filtrate joins the water from the thickening step for recycling.

Incineration. The sludge paste from filtration is passed to a methano-
fueled rotary incinerator having a capacity of 7 t/h. The ash obtained
(at the rate of about 30 t/day) is deposited in a special waste dump.

Process control

Control of the process is provided by automatic measurement and recording of control parameters throughout the operation, automatic control of the process steps and supervision and intervention by the operator from a central control panel.

Corrosion problems

To prevent corrosion due to acids in the waste water and the added reagents (e.g. ferrous sulphate and sulphuric acid), the parts of the plant subject to such effects must be protected as follows:

- (a) Epoxy-resin coatings should be applied to the entire primary treatment stage;
- (b) Acid-resistant brick lining should be provided for the primary reagent dissolving unit, the reagent preparation and feeding unit, the ferrous sulphate tanks and the process channels;
- (c) The stirring devices in the ferrous sulphate tanks must be constructed of stainless steel.

Annex 1

LIMIT VALUES FOR AROMATIC MATERIALS
IN THE WATER SUPPLY

No.	Name of material	USA		USSR	
		ppm	mg/m ³	A/	ppm/m ³
1	2	3	4	5	6
1.	Aceto-propyl acetate	-	-		5
2.	Amyl acetate	-	-		100
3.	Amyl acetate (n)	100	525		-
4.	Amyl acetate (sec)	125	650		-
5.	Benzoacetopropyl acetate	-	-		0,5
6.	Butyl acetate	-	-		200
7.	Butyl acetate (n)	150	710		-
8.	Butyl acetate (sec)	200	950		-
9.	Butyl acetate (tert)	200	950		-
10.	Chloroacetopropyl acetate	-	-		2
11.	Styryl acetate	400	1400		200
12.	Ethylglycol acetate	100	540	x	-
13.	Hexyl acetate (sec)	50	300		-
14.	Isobutyl acetate	100	525		-
15.	Isobutyl acetate	150	700		-
16.	Isopropyl acetate	250	950		-
17.	Methyl acetate	200	610		100
18.	Methylglycol acetate	25	120	x	-

a/ Letters in this column give additional information about the noxious materials or their United States limit values, as follows:

x = Striking toxicity

P = Threshold value

C = Carcinogenic

CE = Carcinogenic experimental

b/ The letter x in this column means that in the USSR standards, the material is said to have striking toxicity.

1	2	3	4	5	6	7
19.	Propyl acetate (n)	200	200	-	-	-
20.	Propyl acetate	-	-	200	-	-
21.	Vinyl acetate	10	10	10	-	-
22.	Acetanilid	-	-	-	-	-
23.	2,3,4 Triethyl-1,4 dihydroquinoline	-	-	-	-	-
24.	Acetone	1000	2400	200	-	-
25.	Acetone cyanohydrin	-	-	0.9	-	x
26.	Acetonitrile	40	70	10	-	-
27.	Acetophenone	-	-	5	-	x
28.	Isopropyl acetoxy-carbamate (n)	-	-	2	-	-
29.	2-Acetylaminofluorene	-	-	xCE	-	-
30.	Acetic acid	10	25	5	-	-
31.	Acrylic acid	-	-	5	-	-
32.	Amino-ocnantic acid	-	-	8	-	-
33.	Aminopelargonic acid	-	-	8	-	-
34.	Boric acid	-	-	10	-	-
35.	Hydrobromic acid	3	10	-	-	-
36.	Butyric acid	-	-	10	-	-
37.	Caproic acid	-	-	5	-	-
38.	Hydrochloric acid	5	7	P	5	-
39.	4-Chlorobenzophenon-2-carboxylic acid	-	-	1	-	-
40.	Chloropelargonic acid	-	-	5	-	-
41.	Chloropropionic acid	-	-	5	-	-
42.	Hydrocyanic acid	10	11	x	0.3	x
43.	Hydrofluoric acid	3	2	0.5	-	-
44.	Formic acid	5	9	-	-	-
45.	Methacrylic acid	-	-	10	-	-
46.	Nitric acid	2	5	-	-	-
47.	Oxalic acid	-	1	-	-	-

1	2	3	4	5	6
47.	Maleic acid	-	-		2
48.	Maleic anhydride	-	1		-
49.	Maleic acid	-	0.1	x	-
51.	Maleic acid	-	1		2
52.	Tartronic acid	-	-		0.1
53.	Maleic furolic acid	-	-		0.1
54.	Tri lauric acid	-	-		2
55.	Trilauric acid	-	-		5
56.	Amylamine	0.1	0.25		0.7
57.	Acrylonitrile	-	0.3	x	-
58.	Butyl acrylate	-	-		10
59.	Ethyl acrylate	25	100	x	-
60.	Methyl acrylate	10	35	x	20
61.	Acrylonitrile	20	45	x	0.5
62.	Piperazine adipat	-	-		5
63.	Adipodinitrile	-	-		20
64.	Acetopropyl alcohol	-	-		10
65.	Allylic alcohol	2	5	x	2
66.	Amylic alcohol	-	-		10
67.	Butyl alcohol	100	300		10
68.	Butyl alcohol (sec)	150	450		-
69.	Butyl alcohol (tert)	100	300		-
70.	2-Chloroethyl alcohol				
71.	Ethylene chlorhydrin				
72.	Crotonic alcohol	-	-		2
73.	Decyl alcohol (n)	-	-		10
74.	1,3-dichlorisopropyl alcohol				
75.	Dichlorohydrin				
76.	Ethyl alcohol	1000	1900		1000

1	2	3	4	5	6	7
77.	Furfuryl alcohol	50	200		-	
78.	heptylic alcohol (n)	-	-		10	
79.	Hexyl alcohol (n)	-	-		10	
80.	Isomyl alcohol	100	360		-	
81.	Isotutyl alcohol	100	300		-	
82.	Isooctyl alcohol	-	-		50	
83.	Isopropyl alcohol	400	950		-	
84.	Methylamylic alcohol	25	100	x	-	
85.	Methanol	200	260		5	x
86.	Nonylalcohol (n)	-	-		10	
87.	Octafluoramylic alcohol	-	-		20	
88.	Octylalcohol (n)	-	-		10	
89.	Propargyl alcohol	1	-	x	1	
90.	Propyl alcohol	200	500		10	
91.	Tetrafluorpropylic alcohol	-	-		20	
92.	Trifluorbutyl alcohol	-	-		20	
93.	Trifluoroethyl alcohol	-	-		10	
94.	Unsaturated alcohols from the fatty alcohol series (allyl, crotonic)	-	-		2	
95.	Acetic aldehyde	200	300		5	
96.	Butyraldehyde	-	-		5	
97.	Chloroacetic aldehyde	1	3	P	-	
98.	Crotonic aldehyde	2	6		0.5	
99.	2-ethylhexaldehyde	-	-		3	
100.	Formaldehyde	2	3	P	0.5	
101.	Furfuryl aldehyde	5	20	x	10	
102.	Isobuthyl aldehyde	-	-		5	
103.	Propionic aldehyde	-	-		5	
104.	Aldrin	-	0.25	x	0.01	x

1	2	3	4	5	6	7
105.	Aluminium (oxide)					
106.	Aluminium, cerium					
107.	Ammonia chromate alum					
108.	Carbonyl (ammonium sulphate)					
109.	Alumina	-	10			-
110.	Starch	-	10			-
111.	Primary aliphatic amines (C ₇ -C ₉)	-	-			1
112.	Secondary aliphatic amines (C ₁₅ -C ₁₉)	-	-			1
113.	Alpha-aminoanthraquinone	-	-			5
114.	4-Aminodiphenyl	-	-	xC		-
115.	2-Aminoethanol					
116.	Ethanolamine					
117.	5-Amino-3-oxo-3,7-dihydro-1,4-naphthoquinonidine	-	-			1
118.	Aminoplastics (powders for moulding)	-	-			6
119.	2-Aminopyridine	0.5	2			-
120.	Aminopyrimidine					
121.	2-Methyl-4-amino-5-ethoxy-methylpyrimidine					
122.	Amino alpha-trifluorotoluene		-			0.5
123.	Ammonia	50	35		20	
124.	Ammonium (sulphamate)					
125.	Ammonium sulphamate					
126.	Anabasine (sulphate)	-	-			0.1
127.	Acetic anhydride	5	20			-
128.	Arsenious anhydride					
129.	Arsenic (arsenious anhydride)					
130.	Arsenic anhydride					
131.	Arsenic (arsenic anhydride)					
132.	Boric anhydride					

1	2	3	4	5	6	7
133.	Boron (oxide)					
134.	Butyl anhydride	-	-		1	
135.	Carbonic anhydride	5000	9000		-	
136.	Chromic anhydride					
137.	Chrome (oxide hexavalent)					
138.	Maleic anhydride	0.25	1		1	
139.	Methacrylic anhydride	-	-		1	
140.	Phthalic anhydride	2	12		1	
141.	Phosphoric anhydride					
142.	Phosphorus (pentoxide)					
143.	Selenious anhydride					
144.	Selenium(dioxide)					
145.	Sulphurous anhydride					
146.	Sulphur (dioxide)					
147.	Sulphuric anhydride					
148.	Sulphur (trioxide)					
149.	Aniline	5	19	x	0.1	x
150.	Anisidine (p)	-	-		1	x
151.	9,10-Anthraquinone	-	-		5	
152.	Anisidine (o- and p-)	-	0.5	x	-	
153.	Stibium and compounds (Sb)	-	0.5		-	
154.	Metallic stibium (under powder form)	-	-		0.5	
155.	Stibium (fluorides and chlorides III and IV expressed as Sb)	-	-		0.3	
156.	Stibium hydride					
157.	Hydrogen stibiate					
158.	Stibium (oxides and sulphides III powders, expressed in Sb)					
159.	Stibium (oxides and sulphides IV powders, expressed in Sb)					
160.	Oxygenated water - Hydrogen peroxide					

1	2	3	4	5	6
161.	Silver (not in soluble compounds)	-	0.01	-	-
162.	Arsenic and its compounds (expressed as As, excepting hydrogen arsenate)	-	0.5	-	-
163.	Arsenic (arsenious and arsenic anhydride)	-	-	-	0.3
164.	Arsine	-	-	-	-
	Arsenized hydrogen				
165.	Nitrogen (nitric oxide)	25	30	-	-
166.	Nitrogen (peroxide)	5	9	P	-
167.	Nitrogen (oxides) expressed as NO ₂	-	-	-	5
168.	Nitrogen (trifluoride)	10	29	-	-
169.	Isopropyl nitrate	-	-	-	5
170.	Propyl nitrate (n)	25	110	-	-
171.	Isopropyl nitrate	-	-	-	1
172.	Barium (soluble compounds)	-	0.5	-	-
173.	Benzene	25	30	Px	5
174.	Benzidine and salts	-	-	C	-
175.	Fuel gasoline (bitumen, cracking, etc. expressed in C)	-	-	-	100
176.	Solvent gasoline (expressed in C)	-	-	-	300
177.	Benzoquinone (p-)	0.1	0.4	-	0.05
178.	Berilium	-	0.002	-	-
179.	Berilium and compounds (Ee)	-	-	-	0.001
180.	Boron (oxide)	-	10	-	5
181.	Boron (tribromide)	1	10	-	-
182.	Boron (trifluoride)	1	3	P	1
183.	Bromine	0.1	0.7	-	-
184.	Bromine (pentafluoride)	0.1	0.7	-	-
185.	Brombenzene	-	-	-	3
186.	Bromchloromethane	200	1050	-	-
187.	1,3-Bromchloropropane	-	-	-	3

1	2	3	4	5	6	7
188.	Bromethane	200	390		5	
189.	Bromethylene	250	1100		-	
190.	Bromoform					
191.	Tribromomethane					
192.	Bromomethane	-	-		1	
193.	Bromotrifluoroethane	1000	6100		-	
194.	Ethyl bromide					
195.	Bromethane					
196.	Ethylene bromide					
197.	1,2 Dibromomethane					
198.	Methyl bromide					
199.	Bromomethane					
200.	Vinyl bromide					
201.	Bromethylene					
202.	1,3-Butadiene	1000	2200		100	
203.	2-Butanone					
204.	Methylethylketone					
205.	2-Butene	-	-		100	
206.	2-Butoxyethanol					
	Butylglycol					
207.	Butylamine	5	15		Px 10	
208.	Butylglycol	50	240		x -	
209.	Butylmercaptan	0.5	1.5		-	
210.	2-Butylthiobenzothiazole	-	-		10	
211.	Butyltoluene (p-tert)	10	60		-	
212.	1,4 Butendiol	-	-		1	
213.	Methyl butyrate	-	-		5	
214.	Cadmium (oxide) (vapours in Cd)	-	0.1	P	0.1	

1	2	3	4	5	6	7
215.	Cadmium (metallic powders and soluble salts)	-	0.2	-		
216.	Cadmium (stearate expressed in Cd)	-	-	0.1		
217.	Calcite	-	10	-		
218.	Calcium (arseniate)	-	1	-		
219.	Calcium (carbonate)	-	10	-		
220.	Calcium (oxide)	-	5	-		
221.	Calcium (sulphate)					
222.	Gypsum and plaster for modelling					
223.	Chlorinated camphene	-	0.5	x	-	
224.	Camphor	-	-		3	
225.	Camphor (synthetic)	2	12	-		
226.	Caprolactame	-	-		10	
227.	Carbaryl	-	5		1	
228.	Cyclomexylamine carbonate	-	-		10	
229.	Celulose (paper fibre)	-	10	-		
230.	Ketene	0.5	0.9	-		
231.	Chlordan	-	0.5	x	0.01	x
232.	Chlorine	1	3		1	
233.	Chlorine (dioxine)	0.1	0.3		0.1	
234.	Chlorine (trifluoride)	0.1	0.4	P	-	
235.	Methyl chloroacetate	-	-		5	
236.	Alpha-chloroacetophenone	0.05	0.3	-		
237.	Chloraniline (m-)	-	-		0.05	x
238.	Chloraniline (p-)	-	-		0.3	x
239.	Chlorobenzene (mono)	75	350		50	x
240.	Chlorobenzal malononitrile (o-)	0.05	0.4	x	-	
241.	2-Chloro-4,6-bis(diethylamino)-1,3,5 triazine	-	-		2	
242.	2-Chloro-1,3-butadiene	25	90	x	2	

1	2	3	4	5	6	7
243.	Isopropyl chlorocarbonate	-	-		0.1	
244.	Chlorocyclonexane	-	-		50	
245.	1-Chloro-2,5-epoxypropane					
246.	Epichlorhydrin					
247.	Chlorethane	1000	2600		50	
248.	Chlorethylene	200	770		30	
249.	Chlorphenylene				2	
250.	Chloroform					
	Trichloromethane					
251.	Chloromethane	100	210		5	
252.	5-Chloromethyl-1-furan-carboxylate butyl	-	-		0.5	
253.	Chloroacetophenone					
254.	Chloromethyltrichlorosilan	-	-		1	
255.	Chloronitrobenzene (p-)	-	1	x	-	
256.	Chloronitrobenzene and derivatives	-	-		1	x
257.	1-Chloro-1-nitropropane	20	100		-	
258.	Chlorophenol (p-)	-	-		1	x
259.	3-Chlorophenylcarbamate isopropyl (n)	-	-		2	
260.	Chloropicrin	0.1	0.7		-	
261.	Chloroprene					
	2-Chloro-1,3-butadiene					
262.	3-Chloropropene	1	3		-	
263.	Chlorostyrene	-	-		50	
264.	2-Chlorethyl chlorosulphamate	-	-		0.3	x
265.	Chlorotene (chlorinated dycycle compounds)	-	-		0.2	
266.	Diethyl chlorothiophosphate	-	-		1	
267.	Dimethyl chlorothiophosphate	-	-		0.5	
268.	Alpha-chlorotoluene	1	5		0.5	

1	2	3	4	5	6	7
269.	Chloroacetylchloropropane	-	-		1.	x
270.	Acryloyl chloride	-	-		0.3	
271.	Allyl chloride					
272.	3-Chloropropane					
273.	Ammonium chloride (fume)	-	10		-	
274.	Benzoyl chloride					
275.	Alpha-trichlorotoluene					
276.	Benzoyl chloride	-	-		5	
277.	Benzyl chloride					
278.	Alpha-chlorotoluene					
279.	Benzal chloride					
280.	Alpha-dichlorotoluene					
281.	Cyan chloride	-	-		0.1	
282.	5-ethoxy-phenyl-1,2-thiazotiazine chloride	-	-		0.2	
283.	Ethyl chloride					
284.	Chloroethane					
285.	Ethylene chloride					
286.	1,2-Dichloroethane					
287.	Ethyliden chloride					
288.	1,1-Dichloroethane					
289.	Isobutylene chloride					
290.	1,2-Dichloro-2-methylpropane					
291.	Methacryloyl chloride	-	-		0.3	
292.	Methyl chloride					
293.	Chloromethane					
294.	Methylene chloride					
295.	Dichloromethane					
296.	Phenacyl chloride					
297.	Alpha-chloroacetophenone					

1	2	3	4	5	6	7
298.	Polyvinyl chloride	-	-		6	
299.	Propylene chloride					
300.	1,2-Dichloropropane					
301.	Trichloroacetyl chloride	-	-		0.1	
302.	Vinyl chloride					
303.	Chlorethylene					
304.	Tertiary butyl chromate (expressed in CrO_3)	-	0.1	Px	-	
305.	Cyclohexylamine chromate	-	-		2	x
306.	Chromates and dichromates expressed in CrO_3	-	-		0.01	
307.	Chromium (metal and insoluble salts)	-	1		-	
308.	Chromium (some insoluble salts)	-	0.1	C	-	
309.	Chromium (Cr III hexahydrated chloride expressed in CrO_3)	-	-		0.01	
310.	Chromium (hexavalent oxide) and chromates (expressed in CrO_3)	-	0.1		-	
311.	Chromium (hexavalent oxide)	-	-		0.01	
312.	Chromium (soluble chromous and chromic salts in Cr)	-	0.5		-	
313.	Chromium (sulphate and ammonium sulphate) in CrO_3	-	-		0.02	
314.	1-methyl-4-diethyl-carbomoyl- piperazine citrate	-	-		5	
315.	Cobalt (vapours and powder)	-	0.1		-	
316.	Cobalt and oxide	-	-		0.5	
317.	Cobalt (dicobalt-orthocarbonyl)	-	-		0.01	
318.	Cobalt hydrocarbonyl and its decomposition products (in CO)	-	-		0.01	
319.	Colophony (decomposition products of the welding bars, expressed in formaldehyde)	-	0.1		-	
320.	Corundum (Al_2O_3)	-	10		-	
321.	Cumaphene	-	0.1		-	

1	2	3	4	5	6	7
322.	Cresols (all isomers)	5	2.	x	-	
323.	Copper	-	-		1	
324.	Copper (vapour)	-	0.1		-	
325.	Copper (powder and mist)	-	1		-	
326.	Copper (trichlorophenate)	-	-		0.1	
327.	Cumene	50	24	x	50	
328.	Free cyanamide	-	-		0.5	x
329.	Allyl cyanate	-	-		0.3	x
330.	Methyl 2-cyanacrylate	2	8		-	
331.	Cyanogen	10	-		-	
332.	Cyanides (expressed in CN)	-	5	x	-	
333.	Cyanides (expressed in HCN)	-	-		0.3	x
334.	Benzyl cyanide					
335.	Phenylacetonitrile					
336.	Vinyl cyanide					
337.	Acrylonitrile					
338.	Cyclohexane	300	1050		80	
339.	Cyclohexanol	50	200		-	
340.	Cyclohexanone	50	200		10	
341.	Cyclohexanone-oxime	-	-		10	
342.	Cyclohexene	300	1015		-	
343.	Cyclohexylamine	-	-		1	
344.	Cyclopentadiene	75	200		5	
345.	2,4-D	-	10		-	
346.	2,4-D (butyl-ester)	-	-		0.5	
347.	2,4-D (alpha-chloro-crotylic ester)	-	-		1	
348.	2,4-D (octyl ester)	-	-		1	
349.	2,4 (ammonium salt)	-	-		1	
350.	DDT					

1	2	3	4	5	6	7
351.	Zelidan					
352.	Decaboran	0.05	0.3	x	-	
353.	Decaline	-	-		100	
354.	Demeton (mixture of O and S)	-	0.1	x	0.02	x
355.	Demeton-methyl (mixture of O and S)	-	0.5	x	0.1	x
356.	Ethylenglycon diacetate	-	-		30	
357.	Diacetonalcokol	50	240		-	
358.	1,2 Diaminoethane	10	25		2	
359.	Diazonone	-	0.1	x	-	
360.	Diazomethane	0.2	0.4		-	
361.	Diborane	0.1	0.1		-	
362.	Dibromchlorotrifluorethane	-	-		50	
363.	Dibromdofluormethane	100	860		-	
364.	1,2 Dibromethane	25	190	x	-	
365.	Di bromomethane	-	-		10	
366.	1,2-Dibromopropane	-	-		5	
367.	Dibutylaminoethanol (2-N-)	2	14	x	-	
368.	Dichloroacetylene	0.1	0.4	P	-	
369.	3,4-Dichloroaniline	-	-		0.5	
370.	Dichloromethylbenzene	-	-		1	
371.	Dichloromethylnaphathalene	-	-		0.5	
372.	3,3-Chloromethyloxacyclohutane	-	-		0.5	
373.	Dichloromethylxylene	-	-		1	
374.	Dichlorobenzene (o-)	50	300	P	-	
375.	Dichlorobenzene (p-)	75	450		-	
376.	Dichlorobenzene	-	-		20	x
377.	3,3-Dichlorobenzidine	-	-	xCl	-	
378.	1,3-Dichloro-2-butene	-	-		-	1
379.	Dichlorodofluoromethane	1000	4950		-	

1	2	3	4	5	6
380.	1,4-Dichloro-2,5-dimethylhydantoin	-	0.2	-	-
381.	1,1-Dichloroethane	100	400	-	-
382.	1,2-Dichloroethane	50	200	-	-
383.	Dichloroethane	-	-	-	-
384.	1,1-Dichloroethane	-	-	10	x
385.	1,2-Dichloroethane	200	790	50	-
386.	Dichloroethane	1000	4200	-	-
387.	Dichloroethane	-	-	-	-
388.	1,2-Dichloropropane	-	-	5	-
389.	1,3-Dichloropropane	-	-	20	-
390.	3,3-Dichloropropene (synthetic isomer)	-	-	0.5	-
391.	Dichloropropane	500	1740	0.5	-
392.	Dichloroacetone	-	-	50	-
393.	1,1,3-Trichloroacetone	-	-	-	-
394.	1,2-Dichloro-2-methyl-propane	-	-	0.3	-
395.	2,3-Dichloro-1,4-naphthaquinone	-	-	0.5	-
396.	3,4-Dichloronitrobenzene	-	-	1	x
397.	1,1-Dichloro-1-nitroethane	10	60	P	-
398.	Dichloro-octafluorocyclohexane	-	-	1	-
399.	Dichlorophenyl dichlorosilane	-	-	1	-
400.	1,2-Dichloropropane	75	350	10	-
401.	1,3-Dichloropropene	-	-	5	-
402.	2,3-Dichloropropene	-	-	3	-
403.	Dichlorostyrene	-	-	50	-
404.	Dichlorotetrafluoroethane	1000	7000	-	-
405.	Alpha-dichlorotoluene	-	-	0.5	-
406.	Alpha, Alpha'-dichloroxyliene	-	-	-	-
407.	Di(chloromethyl)benzene	-	-	-	-

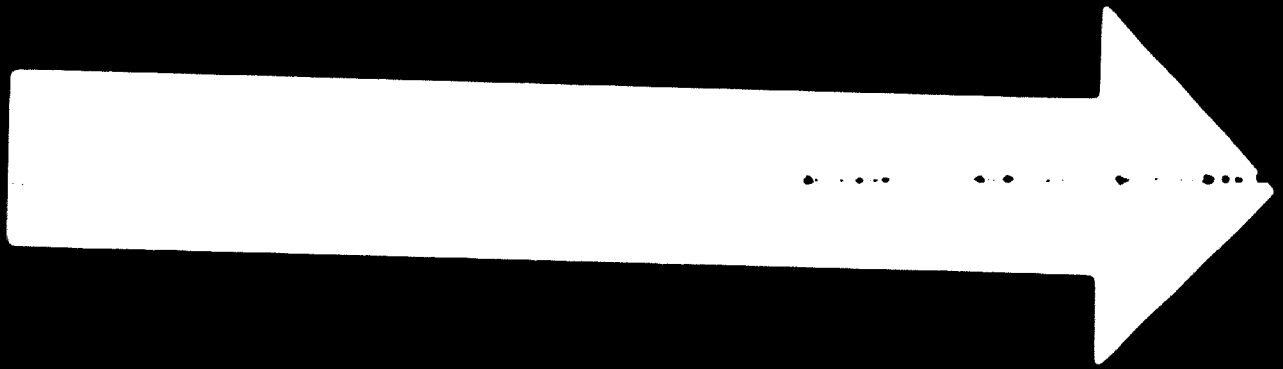
1	2	3	4	5	6	7
408.	Acetylene dichloride					
409.	1,2-Dichloro-ethylene					
410.	Dicumylmethane	-	-		5	x
411.	Dicyclopentadiene	-	-		1	x
412.	Dieldrin	-	0.25	x	0.01	x
413.	Diethylamine	25	75		30	
414.	Diethylaminoethanol	10	50	x	5	
415.	Beta-Diethylaminoethylmercaptan	-	-		1	x
416.	Diethylbenzene	-	-		10	
417.	Diethylenetriamine	1	4	x	-	
418.	Diethylethanolamine					
419.	Diethylamino-ethanol					
420.	Dihydrate perfluoroacetone	-	-		2	x
421.	Diisobutylketone	50	290		-	
422.	4,4'-Diisocyanate diphenylmethane	0.02	0.2	P	-	
423.	Hexamethylene diisocyanate	-	-		0.05	x
424.	Toluylene 2,4-diisocyanate	0.02	0.14	P	-	
425.	Toluylene diisocyanate	-	-		0.5	
426.	Diisopropylamine	5	20	x	5	
427.	Diisopropylbenzene	-	-		50	x
428.	Dimetoxymethane					
429.	Methylal					
430.	Dimethylacetamide	10	35	x	-	
431.	Dimethylamine	10	18		1	
432.	4-Dimethylaminoazobenzene	-	-	CE	-	
433.	3-Dimethylamino-2-chloro-phenotiazine (10-chlorohydrate)	-	-		0.3	x
434.	Dimethylaminoethanol	-	-		5	
435.	Dimethylaniline (n)	5	25	x	0.2	x

1	2	3	4	5	6	7
436.	Dimethylurea	-	-		5	
437.	Dimethylamine					
	Acetone					
438.	Dimethylformamide	-	-		10	
439.	Dimethylmethanamine					
440.	Dimethylaminoethanol					
441.	Dimethylformamide	10	30	x	10	
442.	1,1-Dimethylhydrazine	0.5	1	x	-	
443.	Dinitrobenzene (all isomers)	-	1	x	-	
444.	Dinitrobenzene	-	-		1	x
445.	Dinitroethylphenol (sec)	-	-		0.05	x
446.	Dinitrobenzene (o-)	-	0.2	x	0.05	x
447.	4,6-Dinitro-2-isopropylphenol	-	-		0.05	
448.	Dinitrophenol	-	-		0.05	
449.	Dinitrothiocyanobenzene	-	-		2	x
450.	Dinitrotoluene	-	1.5	x	1	x
451.	1,4-Dioxane	100	360	x	-	
452.	Dioxane	-	-		10	
453.	1,3-Dioxolane	-	-		50	x
454.	Diphenylamine	-	10		-	
455.	Diphenylolpropane	-	-		5	
456.	Dipropylamine (n-)	-	-		2	
457.	Allyl and propyl disulphide	2	12		-	
458.	Dithiometon	-	-		0.1	x
459.	Divinyl					
460.	1,3 Butadiene					
461.	Dodecylmercaptan (tert)	-	-		5	
462.	Emery	-	10		-	
463.	Endosulphane	-	0.1	x	0.1	x

1	2	3	4	5	6	7
466.	Endrin	-	0.1	x	-	
465.	Epichlorohydrine	5	19	x	1	
467.	EPN	-	0.5	x	-	
467.	2,3-Epoxy-1-propanol Glycol					
468.	EPTC	-	-		2	
469.	Tin (dioxide)	-	10		-	
470.	Tin (inorganic compounds)	-	2		-	
471.	Tin (organic compounds)	-	0.1	x	-	
472.	Ethanolamine	3	6		-	
473.	Ether					
474.	Oxide					
475.	Beta-ethoxypropionitrile	-	-		50	
476.	Ethylamine	10	18		-	
477.	Ethylamylketone (sec)	25	130		-	
478.	Ethylbenzene	100	435		-	
479.	Ethylbutylketone	50	230		-	
480.	Ethylenechlorohydrine	5	16	x	0.5	x
481.	Ethylene cyanhydrine	-	-		10	
482.	Ethylene diamine					
483.	1,2 Diaminoethane					
484.	Ethylene imine	0.5	1	xCE	0.02	x
485.	Ethylglycol	200	740	x	-	
486.	2-Ethylhexenal					
487.	2-Ethyl-hexenoic aldehyde					
488.	Ethylmercaptan	0.5	1		1	
489.	Ethylmorpholine (n-)	20	94	x	5	
490.	Ethyltoluene	-	-		50	
491.	Ethoxy aniline	-	-		0.2	x

1	2	3	4	5	6	7
485.	Iron (II) chloride (anhydrous)	-	-	-	0.5	
492.	Iron (II) chloride (hexahydrate)	5	10	-	5	x
494.	Iron (II) chloride	-	-	-	6	
495.	Iron (II) chloride	-	-	x	-	
496.	Iron (II) chloride	-	-	-	0.15	
497.	Iron (II) chloride	-	-	-	0.3	x
498.	Iron (II) chloride	-	-	-	2	
499.	Iron (II) chloride	-	0.1	x	-	
500.	Iron (III) chloride	5	10	x	-	
501.	Iron (II) chloride	-	-	-	1	
502.	Iron (II) chloride (vapor)	-	10	-	-	
503.	Iron (II) chloride (red oxide)	-	10	-	-	
504.	Iron (II) chloride (salts in Fe)	-	1	-	-	
505.	Iron pentacarbonyl	0.01	0.03	-	-	
506.	Iron	-	10	-	-	
507.	Ferrovandium (powder)	-	1	-	-	
508.	Ferrovandium	-	-	-	1	
509.	Fluorine	0.1	0.2	-	-	
510.	Fluorides (expressed in F)	-	2.5	-	-	
511.	Fluorides (expressed in HF)	-	-	-	1	
512.	Benzenyl fluoride					
513.	Alpha-trifluorotoluene					
514.	Benzenyl-m-nitro fluoride					
515.	Alpha-trifluor-m-nitrotoluene					
516.	Fluorizirconate	-	-	-	1	
517.	Formalglycol					
518.	1,3 Dioxalane					
519.	Formamide	-	-	-	3	
520.	Ethyl formate	100	300	-	-	

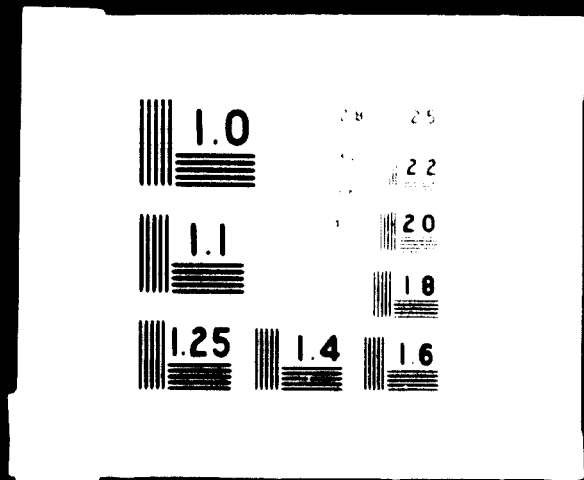
1	2	3	4	5	6	7
521.	Butyl formate	100	250	-	-	-
522.	Formalin	-	-	-	-	-
523.	Formaldehyde	-	-	-	-	-
524.	Phosgene	0.1	0.4	0.5	-	-
525.	Dibutyl phosphate	1	5	-	-	-
526.	Phenyl and diphenyl phosphate (2-ethyl-hexyl)	-	-	1	-	-
527.	Tributyl phosphate	-	5	0.5	-	x
528.	Tricresyl phosphate (o-)	-	0.1	-	-	-
529.	Tricresyl phosphate (containing less 5% ortho-isomers)	-	-	0.5	-	x
530.	tricresyl phosphate (containing more than 5% ortho-isomers)	-	-	0.1	-	x
531.	Triphenyl phosphate	-	3	-	-	-
532.	Trixylenyl phosphate	-	-	1.5	-	x
533.	Phosphine	-	-	-	-	-
534.	Hydrogen phosphide	-	-	-	-	-
535.	Yellow phosphorus	-	0.1	0.03	-	-
536.	Phosphorus (pentachloride)	-	1	-	-	-
537.	Phosphorus (pentasulphide)	-	1	-	-	-
538.	Phosphorus (pentoxide)	-	-	1	-	-
539.	Phosphorus (trichloride)	0.5	3	-	-	-
540.	Dibutyl phthalate	-	5	0.5	-	-
541.	Dimethyl phthalate	-	5	-	-	-
542.	Diethyl phthalate (sec)	-	5	-	-	-
543.	Furan	-	-	0.5	-	-
544.	Butyl 2-furancarboxilate	-	-	0.5	-	-
545.	Furfural	-	-	-	-	-
546.	Furfurylic aldehyde	-	-	-	-	-
547.	Germanium (oxide)	-	-	2	-	-



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1	2	3	4	5	6	7
548.	Germanium (tetrachloride) expressed in Ge)	-	-		1	
549.	Glycerine	-	10		-	
550.	Glycidol	50	150		-	
551.	Graphite (synthetic)	-	10		-	
552.	Gypsum	-	10		-	
553.	Hafnium	-	0.5		-	
554.	HCH	-	-		0.1	x
555.	Gamma-HCH	-	0.5		0.05	x
556.	Heptachlor	-	0.5	x	0.01	x
557.	Heptane (n)	500	2000		-	
558.	Hexachloracetone	-	-		0.5	
559.	Hexachlorobenzene	-	-		0.9	x
560.	Hexachlorocyclopentadione	-	-		0.01	x
561.	Hexachlorethane	1	10	x	-	
562.	Hexachloronaphthalene	-	0.2	x	-	
563.	Hexafluoropropylene	-	-		5	
564.	Hexamethylenediamine	-	-		1	
565.	Hexamethylenimine	-	-		0.5	
566.	Hexamethylenethiocarbamate S-ethyl (N,N-)					
567.	Molinate					
568.	Hexane	500	1800		-	
569.	Hexogene	-	1.5	x	1	
570.	Hexone					
571.	Methylisobutylketone					
572.	Hydrazine	1	1.3	x	-	
573.	Hydrazine, hydrate and derivatives	-	-		0.1	x
574.	Saturated aliphatic hydrocarbons C ₁ -C ₁₀ (expressed in C)	-	-		300	

1	2	3	4	5	6	7
575.	Hydrogen stibiate	0.1	0.5		-	
576.	Hydrogen arseniate	0.05	0.2	0.5		x
577.	Hydrogen phosphide	0.3	0.4		0.1	
578.	Hydrogen selenide	0.05	0.2		-	
579.	Hydrogen sulphide	10	15		10	x
580.	Hydrogen sulphide in mixture with hydrocarbons C ₁ -C ₅	-	-		3	
581.	Cumene hydroperoxide	-	-		1	
582.	Hydroquinone	-	2		-	
583.	Hydroxydiphenylamine (p-)	-	-		0.5	
584.	Beta-hydroxyethylmercaptan	-	-		1	x
585.	Beta-hydroxypropionitrile					
586.	Ethylenecyanhydrine					
587.	Indene	10	45		-	
588.	Indium and its compounds (expressed in In)	-	0.1		-	
589.	Iodine	0.1	1	P	1	
590.	Iodomethane	5	28	x	-	
591.	Methyl iodine					
592.	Iodomethane					
593.	Isobutylene	-	-		100	
594.	Methyl isobutirrate	-	-		10	x
595.	Chlorophenyl isocyanate (m- and p-)	-	-		0.5	
596.	3,4-dichloro-phenyl isocyanate	-	-		0.5	
597.	Methyl isocyanate	0.02	0.05	x	0.05	
598.	Isopropene	-	-		40	
599.	Isopropylamine	5	12		1	
600.	Isopropylaminodiphenyl-amine	-	-		2	
601.	Isopropylbenzene					
602.	Cumene					

1	2	3	4	5	6	7
603.	Methyl thiocyanate	-	-		0.1	x
604.	Methyl isovalerianate	-	-		5	x
605.	Keolin	-	10		-	
606.	Lignin (expressed in C)	-	-		300	
607.	Lithium (hydride)	-	0.025		-	
608.	Magnezite	-	10		-	
609.	Magnesium (carbonate)					
610.	Magnezite					
611.	Magnesium (oxide) vapours	-	10		-	
612.	Malathion	-	10	x	0.5	x
613.	Manganese and its compounds (expressed in Mn)	-	5	P	-	
614.	Manganese	-	-		0.3	
615.	Manganese cyclopentadienyl-tricarbonyl	-	-		0.1	
616.	Manganese methylocyclopentadienyltricarbonyl (expressed in Mn)	0.1	0.2	x	-	
617.	Plastic materials with fluorine-4	-	-		10	
618.	MDI					
619.	Diphenylmethane diisocyanate					
620.	Mercury (alkylate compounds)	-	0.01	x	-	
621.	Mercury (all compounds except alkylate compounds)	-	0.05		-	
622.	Mercury	-	-		0.01	
623.	Mercury (mercury chloride)	-	-		0.1	
624.	Mercurethyl (chloride)	-	-		0.005	x
625.	Mercurethyl (phosphate)	-	-		0.005	x
626.	Diethyl-aminoethyl methacrylate	-	-		800	
627.	Methyl methacrylate	100	410		-	
628.	Metoxychlor	-	10		-	
629.	Metoxyphenol (m-)					

1	2	3	4	5	6	7
630.	Methyl oxide and m-hydroxyphenyl					
631.	Methylacetylene					
632.	Propene					
633.	Methylal	1000	3100		-	
634.	Methylamine	10	12		1	
635.	2-Methyl-4-amino-5-ethoxy-methylpyrimidine	-	-		1	
636.	Methylamylketone (n-)	100	465		-	
637.	Methylaniline	2	9	x	-	
638.	2-Methyl-1,3-butadiene					
639.	Isoprene					
640.	Methylbutylketone	100	410		-	
641.	Methylchlorophora					
642.	1,1,1-Trichloroethane					
643.	Methylcyclohexane	500	2000		50	
644.	Methylcyclohexanol	100	470		-	
645.	Methylcyclohexanone (o-)	100	460	x	-	
646.	Methyldihydropyran	-	-		5	x
647.	Methyldipropylenglycol	100	600	x	-	
648.	Methylethylketone	200	590		200	
649.	2-Methyl-5-ethylpyridine	-	-		2	x
650.	Methyl-fluorphenyl-dichlorosilan	-	-		1	
651.	2-Methylfuran					
652.	Sylvan					
653.	Methylglycol	25	80	x	-	
654.	Methylhexylketone	-	-		200	
655.	Methylhydrazine	0.2	0.35	xP	-	
656.	Methylisoamylketone	100	475		-	
657.	Methylisobutylcarbynol					
658.	Methylaminic alcohol					

1	2	3	4	5	6	7
659.	Methylisobutylketone	100	410		-	
660.	Methylacetone	0.5	1		-	
661.	Methylmorpholine (N-)	-	-		5	
662.	1-Methylnaphthalene	-	-		20	
663.	2-Methylnaphthalene	-	-		20	
664.	Methylphenylketone					
665.	Acetophenone					
666.	Methylpropylketone	200	700		200	
667.	Methylpropylglycol	100	360		-	
668.	Methylpyrrolidone	-	-		800	
669.	Alpha-methylstyrene	100	480	P	5	
670.	Methylthiophen (isomers)	-	-		20	
671.	Methylurethane-cyanole-sulpho- hydrazine	-	-			0.05
672.	Methylvinylketone	-	-			0.1
673.	5-Methyl-5-vinylpyridine	-	-			2
674.	6-Methyl-2-vinylpyridine	-	-			0.5
675.	Molybdenum (insoluble compounds)	-	10		6	
676.	Molybdenum (soluble compounds)	-	5		-	
677.	Molybdenum (soluble compounds under the form of condensation aerosols)	-	-			2
678.	Molybdenum (soluble compounds under the form of powders)	-	-			4
679.	Morpholine	20	70	x	0.5	x
680.	Naphthalene	10	50		20	
681.	Chlorinated naphthalene (higher)	-	-		0.5	x
682.	Alpha-naphthoquinone	-	-		0.1	
683.	Beta-naphthylamine	-	-	0	-	
684.	Nickel (metal and soluble compounds expressed in Ni)	-	1		-	
685.	Nickel (oxide and sulphide expressed in Ni)	-	-		0.5	

1	2	3	4	5	6	7
686.	Nickel (salts under the form of aerosols expressed in Ni)	-	-		0.005	
687.	Nickel carbonyl	0.001	0.007 C		0.0005	
688.	Nicotine	-	0.5	x	-	
689.	Nicotine (sulphate)	-	-		0.1	
690.	Nitroaniline (o-)	-	-		0.5	x
691.	Nitroaniline (p-)	1	6	x	0.1	x
692.	Nitroanisol (p-)	-	-		3	
693.	Nitrobenzene	1	5	x	3	x
694.	Nitrobutane	-	-		30	
695.	Nitrochlorbenzene					
696.	Chloronitrobenzene					
697.	Nitrocyclohexane	-	-		1	
698.	4-Nitrodiphenyl	-	-	C	-	
699.	Nitroethane	100	310		30	
700.	Nitroform					
701.	Trinitromethane					
702.	Nitroglycerine	0.2	2	x	-	
703.	Nitromethane	100	250		30	
704.	1-Nitropropane	25	90		30	
705.	2-Nitropropane	25	90		30	
706.	Nitrosodimethylamine (N-)	-	-	xCE	-	
707.	Nitrotoluene	5	30	x	-	
708.	Nitrotrichloromethane					
709.	Chloropicrine					
710.	Nitroxylene	-	-		5	x
711.	Octachloronaphthalene	-	0.1	x	-	
712.	Octane	400	1900		-	
713.	Osmium (tetraoxide)	-	0.002		-	

1	2	3	4	5	6	7
714.	2-Oxohexamethylenimine					
715.	Caprolactame					
716.	Oxycarbamate	-	-		0.5	
717.	Allyl and glucidil oxide	10	45	P	-	
718.	Butyl and glucidil oxide	50	270		-	
719.	Carbon oxide	50	55		20	
720.	2,2'-dichlorodiethyl oxide	10	90	Px	2	x
721.	Dicetyl oxide	400	1200		300	
722.	Diglicidil oxide	0.5	2.8	P	-	
723.	Diphenyl oxide (vapours)	1	7		-	
724.	Diphenyl/biphenyl oxide (vapours)	1	7		-	
725.	Diphenyl chlorate oxide	-	0.5		0.5	-
726.	Ethylene oxide	50	90		1	
727.	Isopropyl and glucidil oxide	50	240		-	
728.	Mesityl oxide	25	100		1	x
729.	Methyl and m-hydroxyphenyl oxide	-	-		0.5	x
730.	Methyl and chloromethyl oxide	-	-		0.5	
731.	Phenyl and glycidil oxide	10	60		-	
732.	Propylene oxide	100	240		1	x
733.	Oxygene (difluoride)	-	-		20	
734.	Ozone	0.05	0.1		-	
735.	Paraformaldehyde					
736.	polyformaldehyde					
737.	Parathion	-	0.1	x	0.05	x
738.	Parathion-methyl	-	0.2	x	0.1	x
739.	Parathion-methyl-ethyl	-	-		0.03	x
740.	Pentaboran	0.005	0.01		-	
741.	Pentachloracetone	-	-		0.5	
742.	Pentachloronaphthalene	-	0.5	x	-	

1	2	3	4	5	6	7
743.	Pentachloronitrobenzene	-	-		0.5	
744.	Pentachlorophenol	-	0.5	x	0.1	x
745.	Pentachlorophenol (Na salt)					
746.	Sodium (pentachlorophenate)					
747.	Pentane	500	1500		-	
748.	Butyl peracetate (tert)	-	-		0.1	
749.	Butyl perbenzoate (tert)	-	-		1	
750.	Perchlorodimethylketone					
751.	Hexachloroacetone					
752.	Perchloroethylene	100	670		10	
753.	Perchloromethylmercaptan	0.1	0.8		1	
754.	Perchloryl (fluoride)	3	14		-	
755.	Diethyl perfluoradipate	-	-		0.1	
756.	Perfluoradiponitrile	-	-		0.1	
757.	Perfluorodimethylketone					
758.	Hexafluoroacetone					
759.	Diethyl perfluorglutarat	-	-		0.1	
760.	Perfluorglutarodinitrile	-	-		0.05	
761.	Perfluoroisobutylene	-	-		0.1	
762.	Dibenzoyl peroxide	-	5		-	
763.	Tertiary dibutyl peroxide	-	-		100	
764.	Hydrogen peroxide	1	1.4		-	
765.	Petroleum (bitumen) (vapours)	-	5		-	
766.	Petroleum (gasoline)					
767.	Naphtha, fuel gasoline					
768.	Petroleum (liquid gas)	1000	1800		-	
769.	Picoline (isomer mixture)	-	-		5	
770.	Chlorinated pipen					
771.	Polychlorpipen					

1	2	3	4	5	6	7
772.	Piperidine	-	-		0.2	x
773.	Pivalone	-	0.1		-	
774.	Platinum (soluble salts in Pt)	-	0.002		-	
775.	Lead and its inorganic compounds	-	-		0.01	
776.	Lead (arsenate)	-	0.15		-	
777.	Lead tetraethyl	-	0.100 x		0.005 x	
778.	Lead tetramethyl	-	0.150 x		-	
779.	Polychloropinen	-	-		0.2	x
780.	High density polyethylene	-	-		10	
781.	Polyformaldehyde	-	-		5	
782.	Polypropylene (unstabilized)	-	-		10	
783.	Potassium (butyl xantogenate)	-	-		10	
784.	Potassium (hydroxide)					
	Caustic alkali					
785.	Propadiene/propin (mixture)	1000	1800		-	
786.	Propanyl	-	-		0.1	
787.	Propazine (10-3(dimethylamino-propyl) phenothiazinchlorhydrate	-	-		5	
788.	Beta-propiolactone	-	-	CE	-	
789.	Methyl propionate	-	-		10	x
790.	Propyl propionate	-	-		70	
791.	Propylamine (N-)	-	-		5	
792.	Propylenimine	2	5	x	-	
793.	Propine	1000	1650		-	
794.	Propine/propadiene (mixture)	1000	1800		-	
795.	Pseudobutilene					
796.	2-Butene					
797.	Rhodium (vapours and metallic powders)	-	0.1		-	
798.	Rhodium (soluble salts)	-	0.001		-	

1	2	3	4	5	6	7
799.	Selenium amorphous	-	-		2	x
800.	Selenium (dioxide)	-	-		0.1	
801.	Selenium (compounds expressed in Se except hydrogen selenide)	-	0.2		-	
802.	Selenium (hexafluoride)	0.05	0.4		-	
803.	Ethyl silicate	100	850		20	
804.	Methyl silicate	5	30	P	-	
805.	Silicon (carbide)	-	10		-	
806.	Sodium (cis-beta-chloroacrylate)	-	-		2	
807.	Sodium (2-(2,4-dichlorophenoxy)ethylsulphate)	-	-		0.5	
808.	Sodium (fluoroacetate)	-	10		-	
809.	Sodium (hydroxide)	-	2		-	
810.	Sodium (pentachlorophenate)	-	-		0.1	x
811.	Sodium (thiocyanate) technical	-	-		50	
812.	Naphtha solvent (expressed in C)	-	-		100	
813.	Glass (fibre 5 - 7 diameter)	-	10		-	
814.	Stibine					
815.	Hydrogen antimonide					
816.	Streptomycin	-	-		0.1	
817.	Stricnine	-	0.15		-	
818.	Styrene	100	420		5	
819.	Styrene and alpha-methyl styrene (copolymer)	-	-		5	
820.	Sulphur (dioxide)	5	13		10	
821.	Sulphur (hexafluoride)	1000	6000		-	
822.	Sulphur (monochloride)	1	6		-	
823.	Sulphur (pentafluoride)	0.025	0.25		-	
824.	Sulphur (trioxide)	-	-		1	
825.	Ammonium sulphamate	-	10		10	

1	2	3	4	5	6	7
826.	Sulphamide	-	-		1	
827.	Dimethyl sulphate	1	5	x	-	
828.	Carbon sulphide	20	60	x	10	
829.	Dimethyl sulphide	-	-		50	x
830.	Ethylene sulphide	-	-		0.1	x
831.	Sulphuryl (fluoride)	5	20		-	
832.	Sylvane	-	-		1	
833.	2,4,5-F	-	10		-	
834.	Tantalum	-	5		-	
835.	Tantalum and oxides	-	-		10	
836.	Tellurium	-	0.1		0.01	
837.	Tellurium (hexafluoride)	0.02	0.2		-	
838.	TEPP	-	0.05	x	-	
839.	Turpentine	100	560		300	
840.	Dimethyl terephthalate	-	-		0.1	
841.	1,1,2,2-Tetrabromethane	1	14		-	
842.	Tetrabromethane	-	-		1	
843.	Acetylene tetrabromide					
844.	1,1,2,-Tetrabromethane					
845.	1,1,2,-Tetrachloro-2,2-difluorethane	500	4170		-	
846.	1,1,2,-Tetrachloro-1,2-difluorethane	500	4170		-	
847.	1,1,2,-Tetrachlorethane	5	25	x	-	
848.	Tetrachloroethane	-	-		5	x
849.	Tetrachlorethylene					
850.	Perchloroethylene					
851.	Tetrachloroheptane	-	-		1	
852.	Tetrachlorhexatriene	-	-		0.3	x
853.	Tetrachloromethane	10	65	x	20	x

1	2	3	4	5	6	7
854.	Tetrachlorononane	-	-		1	
855.	Tetrachloronaphthalene	-	2	x	-	
856.	Tetrachloropentane	-	-		1	
857.	Tetrachloropropane	-	-		1	
858.	Acetylene tetrachloride					
859.	1,1,2,2-Tetrachloroethane					
860.	Carbon tetrachloride					
861.	Tetrachloromethane					
862.	Tetrahydrofuran	200	590		100	
863.	Tetraline	-	-		100	
864.	Tetranitromethane	1	3		0.3	
865.	Thallium (Soluble compounds in T1)	-	0.1	x	-	
866.	Thallium (bromide and iodine)	-	-		0.01	
867.	Thiodiphenylamine					
868.	Phenothiazine					
869.	Thiofuran	-	-		20	
870.	O,O'-dimethyl and O-ethoxycarbonyl-methyl thiophosphate	-	-		1	x
871.	O-methyl, O-ethyl and O-trichloro-phenyl thiophosphate	-	-		0.3	x
872.	Tyram	-	5		0.5	
873.	Thorium	-	-		0.05	
874.	Titanium (dioxide)	-	10		-	
875.	Titanium and dioxide	-	-		10	
876.	Titanium (tetrachloride)	-	-		1	
877.	Toluene	-	-		50	
878.	Toluidine (o-)	5	22		-	
879.	Toluidine	-	-		3	x
880.	Toluene diamine	-	-		2	x
881.	Toxaphen					
882.	Chlorinated camphene					

1	2	3	4	5	6	7
883.	Tribromomethane	0.5	5	x	5	
884.	1,1,3-Trichloroacetone	-	-		0.3	
885.	Trichlorobenzene	-	-		10	
886.	Trichloroethyl ketone					
887.	1,1,3-Trichloroacetone					
888.	1,1, Trichloroethane	350	1900		20	
889.	1,1,2-Trichloroethane	10	45	x	-	
890.	Trichloroethylene	100	535		10	
891.	Trichlorofluoromethane	1000	5600		-	
892.	Trichloromethane	50	240	P	-	
893.	Trichloromethylmercaptan					
894.	Perchloromethyl-mercaptan					
895.	Trichloronaphthalene	-	5	x	1	x
896.	Trichlorophenol (Cu salt)					
897.	Cu (Cu trichlorophenate)					
898.	1,2,3,Trichloropropane	50	300		-	
899.	Trichloropropane	-	-		2	
900.	Trichloropropylene	-	-		3	
901.	Trichlorosylane	-	-		1	
902.	Trichlorothiophenol (trichlorothiophenol disulphide and paraffin mixture)-				5	
903.	Trichlorotoluene	-	-		0.2	
904.	Trichlorotriazone					
905.	Cyan chloride					
906.	1,1,2-Trichloro-1,2,2,-trifluoroethane	1000	7600		-	
907.	Triethylamine	25	100		10	
908.	Triethoxysilan	-	-		1	
909.	Trifluoroethylamine	-	-		100	
910.	2-Trifluormethyl-10-3-4-methyl-1-piperidinyl)propyl/phenothiazine dichlorohydrate	-	-			0.01

1	2	3	4	5	6	7
911.	Alpha-trifluor-m-nitrotoluene	-	-		1	
912.	Trifluorpropylamine	-	-		5	
913.	Alpha-trifluortoluene	-	-		100	
914.	Trimethylamine	-	-		5	
915.	Trimethylbenzene	25	120		-	
916.	2,2,4-Trimethyl-1,2-dihydro-quinoline	-	-		1	
917.	Trimethylene trinitramine					
918.	Hexogen					
919.	Trimethylolpropane	-	-		50	
920.	Trinitromethane	-	-		0.5	
921.	2,4,6,-Trinitrophenyl-methyl-nitramine					
922.	Tetryl					
923.	Trinitrotoluene	-	1.5	x	1	x
924.	Tripropylamine	-	-		2	
925.	Trithiophosphate S,S,S-tributyl	-	-		0.2	
926.	Tungsten (insoluble compounds expressed in W)	-	-		-	
927.	Tungsten and carbides	-	-		6	
928.	Tungsten (soluble compounds expressed in W)	-	1		-	
929.	Natural uranium and soluble and insoluble compounds expressed in U)	-	0.2		-	
930.	Uranium (insoluble compounds)	-	-		0.075	
931.	Uranium (soluble compounds)	-	-		0.015	
932.	Methyl valerianate	-	-		1	x
933.	Vanadium (pentoxide)(vapours)	-	0.05	P	-	
934.	Vanadium (pentoxide)(powders)	-	0.5		-	
935.	Vanadium (trioxide and pentoxide) (powder)	-	-		0.5	
936.	Vanadium (powder)	-	-		4	

1	2	3	4	5	6	7
937.	Vinylacetylene	-	-		20	
938.	2-Vinylpyridine	-	-		0.5	
939.	Vinyltoluene	100	480		50	
940.	White spirit	200	1150		-	
941.	White spirit (expressed in C)	-	-		300	
942.	Xylene	100	435		50	
943.	Xylylene	5	25	x	3	x
944.	Yttrium	-	1		-	
945.	Zinc (chloride) vapours	-	1		-	
946.	Zinc (oxide) vapours	-	5		-	
947.	Zinc (oxide)	-	-		6	
948.	Zirconylfluoride					
949.	Zinc pentachlorothiophenate	-	-		2	
950.	Fluorozirconate					
951.	Zirconium (compounds in Zn)	-	5		-	
952.	Zirconium and its insoluble compounds (silicates, dioxide and carbide)	-	-		6	
953.	Zirconium (nitrate)	-	-		4	

Annex II

QUALITY STANDARDS FOR SURFACE-WATER
SUPPLIES IN ROMANIA

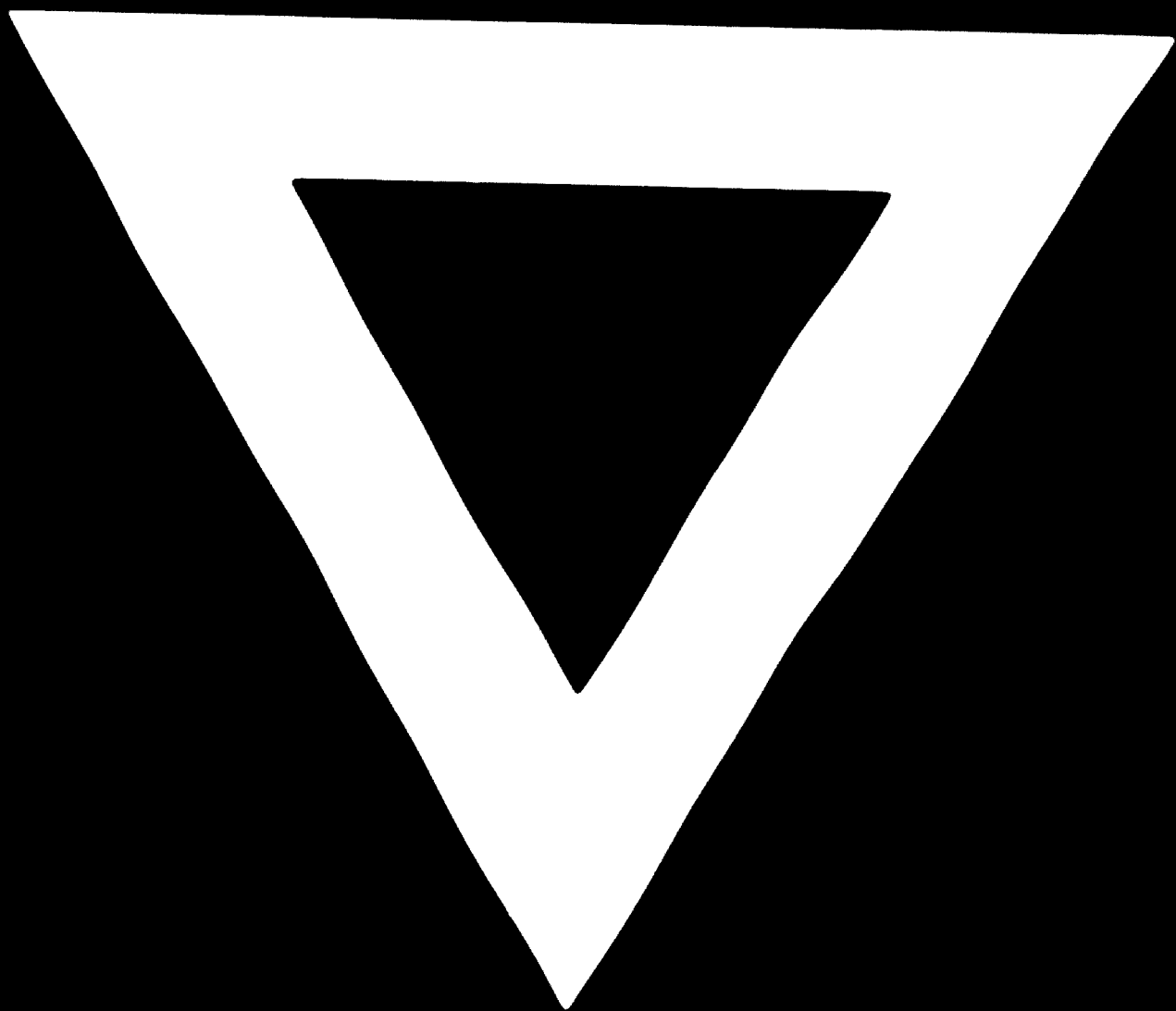
No.	Characteristic	Category ^{a/}		
		I	II	III
1	2	3	4	5
1.	Ammonium (NH ₄), mg/dm ³ max.	1	3	10
2.	Ammonia (NH ₃), mg/dm ³ max.	0.1	0.3	0.5
3.	Silver (Ag), mg/dm ³ max.	0.01	0.01	0.01
4.	Arsenic (As), mg/dm ³ max.	0.05	0.2	0.5
5.	Nitrate (NO ₃), mg/dm ³ max.	10	50	no standard
6.	Nitrate (NO ₂), mg/dm ³ max.	1	3	"
7.	Barium (Ba), mg/dm ³ max.	1	4	4
8.	Benzene, mg/dm ³	0.1	0.1	0.1
9.	Free CO ₂ , mg/dm ³	50	50	50
10.	Cadmium (Cd), mg/dm ³	0.005	0.03	0.2
11.	Calcium (Ca), mg/dm ³	150	200	300
12.	Cyanides (CN), mg/dm ³	0.01	0.2	0.2
13.	Free waste chlorine (Cl ₂), mg/dm ³	0.005	0.005	0.005
14.	Chlorides (Cl), mg/dm ³	250	400	400
15.	Cobalt (Co), mg/dm ³	1	5	5
16.	Hydrogen ion concentration, pH	6.5-8.5	6.5-8.5	6.5-9.0
17.	Chromium trivalent (Cr), mg/dm ³	0.5	0.5	0.5
18.	Chromium hexavalent (Cr), mg/dm ³	0.05	0.05	0.1
19.	Copper (Cu), mg/dm ³	0.1	0.1	3
20.	Active anion detergents, mg/dm ³	0.5	1	3

^{a/} See text (Chapter II) for explanation of categories.

1	2	3	4	5
21.	Phenols carried off by water vapours, mg/dm ³	0.001	0.02	0.05
22.	Total iron (Fe), mg/dm ³	0.3	1	1
23.	Fluorine (F), mg/dm ³	0.5	1	1
24.	Ammonium (N), mg/dm ³	0.1	0.1	0.1
25.	Urea, mg/dm ³	0.5	0.5	0.5
26.	Hydrogen sulphide and sulphides (H ₂ S), mg/dm ³	none		0.1
27.	Magnesium (Mg), mg/dm ³	50	100	200
28.	Manganese (Mn), mg/dm ³	0.1	0.3	0.3
29.	Mercury (Hg), mg/dm ³	0.005	0.01	0.02
30.	Naphthalene, mg/dm ³	0.1	0.1	0.1
31.	Nickel (Ni), mg/dm ³	0.1	0.1	0.1
32.	Oxygen dissolved in water (O ₂), mg/dm ³ min.	6	5	4
33.	Lead (Pb), mg/dm ³	0.05	0.1	0.1
34.	Fix waste, mg/dm ³	750	1000	1200
35.	Selenium (Se), mg/dm ³	0.01	0.01	0.01
36.	Sodium (Na), mg/dm ³	100	200	200
37.	Mercaptanic sulphur, mg/dm ³	0.5	0.5	0.5
38.	Sulphates (SO ₄), mg/dm ³	200	400	400
39.	Carbon sulphide, mg/dm ³	1	1	1
40.	Organic materials			
	a) Biological oxygen consumption (BOD ₅), mg/dm ³	5	7	12
	b) Chemical oxygen consumption (COC), mg/dm ³			
	- COC (Mn)	10	15	25
	- COC (Cr)-with K-dichromate	10	20	30

1	2	3	4	5
41. Tannin and lignin, mg/dm ³		10	10	15
42. Toluene, mg/dm ³		0.1	0.1	0.1
43. Oil products, mg/dm ³		0.1	0.1	0.1
44. Zinc (Zn), mg/dm ³		0.01,	0.1	0.1
45. Physical and organoleptic characteristics				
Radioactivity Ci/cm ³				
- overall alpha		1.10 ⁻⁹	1.10 ⁻⁹	1.10 ⁻⁹
- overall beta		50.10 ⁻⁹	50.10 ⁻⁹	50.10 ⁻⁹
46. Colour		colourless		no standard
47. Smell		without smell		
48. Bacteriologic characteristics				
Coli bacillus/dm ³ max.		100,000		no standard





76. 06. 30